

ANNALS OF THE NEW YORK ACADEMY OF SCIENCES

Issue: *Ecological Economics Reviews***Full cost accounting for the life cycle of coal**

Paul R. Epstein,¹ Jonathan J. Buonocore,² Kevin Eckerle,³ Michael Hendryx,⁴ Benjamin M. Stout III,⁵ Richard Heinberg,⁶ Richard W. Clapp,⁷ Beverly May,⁸ Nancy L. Reinhart,⁸ Melissa M. Ahern,⁹ Samir K. Doshi,¹⁰ and Leslie Glustrom¹¹

¹Center for Health and the Global Environment, Harvard Medical School, Boston, Massachusetts. ²Environmental Science and Risk Management Program, Department of Environmental Health, Harvard School of Public Health, Boston, Massachusetts.

³Accenture, Sustainability Services, Philadelphia, Pennsylvania. ⁴Department of Community Medicine, West Virginia

University, Morgantown, West Virginia. ⁵Wheeling Jesuit University, Wheeling, West Virginia. ⁶Post Carbon Institute, Santa

Rosa, California. ⁷Boston University School of Public Health, Boston, Massachusetts. ⁸Kentuckians for the Commonwealth,

London, Kentucky ⁹Department of Pharmacotherapy, Washington State University, Spokane, Washington. ¹⁰Gund Institute for

Ecological Economics, University of Vermont, Burlington, Vermont. ¹¹Clean Energy Action, Boulder, Colorado

Address for correspondence: Paul R. Epstein, M.D., M.P.H., Center for Health and the Global Environment, Harvard Medical School, Landmark Center, 401 Park Drive, Second Floor, Boston, Massachusetts 02215. paul_epstein@hms.harvard.edu

Each stage in the life cycle of coal—extraction, transport, processing, and combustion—generates a waste stream and carries multiple hazards for health and the environment. These costs are external to the coal industry and are thus often considered “externalities.” We estimate that the life cycle effects of coal and the waste stream generated are costing the U.S. public a third to over one-half of a trillion dollars annually. Many of these so-called externalities are, moreover, cumulative. Accounting for the damages conservatively doubles to triples the price of electricity from coal per kWh generated, making wind, solar, and other forms of nonfossil fuel power generation, along with investments in efficiency and electricity conservation methods, economically competitive. We focus on Appalachia, though coal is mined in other regions of the United States and is burned throughout the world.

Keywords: coal; environmental impacts; human and wildlife health consequences; carbon capture and storage; climate change

Preferred citation: Paul R. Epstein, Jonathan J. Buonocore, Kevin Eckerle, Michael Hendryx, Benjamin M. Stout III, Richard Heinberg, Richard W. Clapp, Beverly May, Nancy L. Reinhart, Melissa M. Ahern, Samir K. Doshi, and Leslie Glustrom. 2011.

Full cost accounting for the life cycle of coal in “Ecological Economics Reviews.” Robert Costanza, Karin Limburg & Ida Kubiszewski, Eds. *Ann. N.Y. Acad. Sci.* 1219: 73–98.

Introduction

Coal is currently the predominant fuel for electricity generation worldwide. In 2005, coal use generated 7,334 TWh (1 terawatt hour = 1 trillion watt-hours, a measure of power) of electricity, which was then 40% of all electricity worldwide. In 2005, coal-derived electricity was responsible for 7.856 Gt of CO₂ emissions or 30% of all worldwide carbon dioxide (CO₂) emissions, and 72% of CO₂ emissions from power generation (one gigaton = one billion tons; one metric ton = 2,204 pounds.)¹ Non-power-generation uses of coal, including industry (e.g., steel, glass-blowing), transport, residential services, and agriculture, were responsible for another 3.124 Gt of CO₂, bringing coal’s total burden of CO₂ emissions to 41% of worldwide CO₂ emissions in 2005.¹

By 2030, electricity demand worldwide is projected to double (from a 2005 baseline) to 35,384 TWh, an annual increase of 2.7%, with the quantity of electricity generated from coal growing 3.1% per annum to 15,796 TWh.¹ In this same time period, worldwide CO₂ emissions are projected to grow 1.8% per year, to 41.905 Gt, with emissions from the coal-power electricity sector projected to grow 2.3% per year to 13.884 Gt.¹

In the United States, coal has produced approximately half of the nation’s electricity since 1995,² and demand for electricity in the United States is projected to grow 1.3% per year from 2005 to 2030, to 5,947 TWh.¹ In this same time period, coal-derived electricity is projected to grow 1.5% per year to 3,148 TWh (assuming no policy changes from the present).¹ Other agencies show similar projections; the U.S. Energy Information Administration (EIA)

projects that U.S. demand for coal power will grow from 1,934 TWh in 2006 to 2,334 TWh in 2030, or 0.8% growth per year.³

To address the impact of coal on the global climate, carbon capture and storage (CCS) has been proposed. The costs of plant construction and the “energy penalty” from CCS, whereby 25–40% more coal would be needed to produce the same amount of energy, would increase the amount of coal mined, transported, processed, and combusted, as well as the waste generated, to produce the same amount of electricity.^{1,4} Construction costs, compression, liquefaction and injection technology, new infrastructure, and the energy penalty would nearly double the costs of electricity generation from coal plants using current combustion technology (see Table 2).⁵

Adequate energy planning requires an accurate assessment of coal reserves. The total recoverable reserves of coal worldwide have been estimated to be approximately 929 billion short tons (one short ton = 2,000 pounds).² Two-thirds of this is found in four countries: U.S. 28%; Russia 19%; China 14%, and India 7%.⁶ In the United States, coal is mined in 25 states.² Much of the new mining in Appalachia is projected to come from mountaintop removal (MTR).²

Box 1.

Peak Coal?

With 268 billion tons of estimated recoverable reserves (ERR) reported by the U.S. Energy Information Administration (EIA), it is often estimated that the United States has “200 years of coal” supply.⁷ However, the EIA has acknowledged that what the EIA terms ERR cannot technically be called “reserves” because they have not been analyzed for profitability of extraction.⁷ As a result, the oft-repeated claim of a “200 year supply” of U.S. coal does not appear to be grounded on thorough analysis of economically recoverable coal supplies.

Reviews of existing coal mine lifespan and economic recoverability reveal serious constraints on existing coal production and numerous constraints facing future coal mine expansion. Depending on the resolution of the geologic, economic, legal, and transportation constraints facing future coal mine expansion, the planning horizon for moving beyond coal may be as short as 20–30 years.^{8–11}

Recent multi-Hubbert cycle analysis estimates global peak coal production for 2011 and U.S. peak coal production for 2015.¹² The potential of “peak coal” thus raises questions for investments in coal-fired plants and CCS.

Worldwide, China is the chief consumer of coal, burning more than the United States, the European Union, and Japan combined. With worldwide demand for electricity, and oil and natural gas insecurities growing, the price of coal on global markets doubled from March 2007 to March 2008: from \$41 to \$85 per ton.¹³ In 2010, it remained in the \$70+/ton range.

Coal burning produces one and a half times the CO₂ emissions of oil combustion and twice that from burning natural gas (for an equal amount of energy produced). The process of converting coal-to-liquid (not addressed in this study) and burning that liquid fuel produces especially high levels of CO₂ emissions.¹³ The waste of energy due to inefficiencies is also enormous. Energy specialist Amory Lovins estimates that after mining, processing, transporting and burning coal, and transmitting the electricity, only about 3% of the energy in the coal is used in incandescent light bulbs.¹⁴

Thus, in the United States in 2005, coal produced 50% of the nation’s electricity but 81% of the CO₂ emissions.¹ For 2030, coal is projected to produce 53% of U.S. power and 85% of the U.S. CO₂ emissions from electricity generation. None of these figures includes the additional life cycle greenhouse gas (GHG) emissions from coal, including methane from coal mines, emissions from coal transport, other GHG emissions (e.g., particulates or black carbon), and carbon and nitrous oxide (N₂O) emissions from land transformation in the case of MTR coal mining.

Coal mining and combustion releases many more chemicals than those responsible for climate forcing. Coal also contains mercury, lead, cadmium, arsenic, manganese, beryllium, chromium, and other toxic, and carcinogenic substances. Coal crushing, processing, and washing releases tons of particulate matter and chemicals on an annual basis and contaminates water, harming community public health and ecological systems.^{15–19} Coal combustion also results in emissions of NO_x, sulfur dioxide (SO₂),

the particulates PM₁₀ and PM_{2.5}, and mercury; all of which negatively affect air quality and public health.^{20–23}

In addition, 70% of rail traffic in the United States is dedicated to shipping coal, and rail transport is associated with accidents and deaths.²⁰ If coal use were to be expanded, land and transport infrastructure would be further stressed.

Summary of methods

Life cycle analysis, examining all stages in using a resource, is central to the full cost accounting needed to guide public policy and private investment. A previous study examined the life cycle stages of oil, but without systematic quantification.²⁴ This paper is intended to advance understanding of the measurable, quantifiable, and qualitative costs of coal.

In order to rigorously examine these different damage endpoints, we examined the many stages in the life cycle of coal, using a framework of environmental externalities, or “hidden costs.” Externalities occur when the activity of one agent affects the well-being of another agent outside of any type of market mechanism—these are often not taken into account in decision making and when they are not accounted for, they can distort the decision-making process and reduce the welfare of society.²⁰ This work strives to derive monetary values for these externalities so that they can be used to inform policy making.

This paper tabulates a wide range of costs associated with the full life cycle of coal, separating those that are quantifiable and monetizable; those that are quantifiable, but difficult to monetize; and those that are qualitative.

A literature review was conducted to consolidate all impacts of coal-generated electricity over its life cycle, monetize and tabulate those that are monetizable, quantify those that are quantifiable, and describe the qualitative impacts. Since there is some uncertainty in the monetization of the damages, low, best, and high estimates are presented. The monetizable impacts found are damages due to climate change; public health damages from NO_x, SO₂, PM_{2.5}, and mercury emissions; fatalities of members of the public due to rail accidents during coal transport; the public health burden in Appalachia associated with coal mining; government subsidies; and lost value of abandoned mine lands. All values

are presented in 2008 US\$. Much of the research we draw upon represented uncertainty by presenting low and/or high estimates in addition to best estimates. Low and high values can indicate both uncertainty in parameters and different assumptions about the parameters that others used to calculate their estimates. Best estimates are not weighted averages, and are derived differently for each category, as explained below.

Climate impacts were monetized using estimates of the social cost of carbon—the valuation of the damages due to emissions of one metric ton of carbon, of \$30/ton of CO₂equivalent (CO₂e),²⁰ with low and high estimates of \$10/ton and \$100/ton. There is uncertainty around the total cost of climate change and its present value, thus uncertainty concerning the social cost of carbon derived from the total costs. To test for sensitivity to the assumptions about the total costs, low and high estimates of the social cost of carbon were used to produce low and high estimates for climate damage, as was done in the 2009 National Research Council (NRC) report on the “Hidden Costs of Energy.”²⁰ To be consistent with the NRC report, this work uses a low value of \$10/ton CO₂e and a high value of \$100/ton CO₂e.

All public health impacts due to mortality were valued using the value of statistical life (VSL). The value most commonly used by the U.S. Environmental Protection Agency (EPA), and used in this paper, is the central estimate of \$6 million 2000 US\$, or \$7.5 million in 2008 US\$.²⁰

Two values for mortality risk from exposure to air pollutants were found and differed due to different concentration-response functions—increases in mortality risk associated with exposure to air pollutants. The values derived using the lower of the two concentration-response functions is our low estimate, and the higher of the two concentration-response functions is our best and high estimate, for reasons explained below. The impacts on cognitive development and cardiovascular disease due to mercury exposure provided low, best, and high estimates, and these are presented here.

Regarding federal subsidies, two different estimates were found. To provide a conservative best estimate, the lower of the two values represents our low and best estimate, and the higher represents our high estimate. For the remaining costs, one point estimate was found in each instance, representing our low, best, and high estimates.

The monetizable impacts were normalized to per kWh of electricity produced, based on EIA estimates of electricity produced from coal, as was done in the NRC report tabulating externalities due to coal.^{2,20} Some values were for all coal mining, not just for the portion emitted due to coal-derived electricity. To correct for this, the derived values were multiplied by the proportion of coal that was used for electrical power, which was approximately 90% in all years analyzed. The additional impacts from nonpower uses of coal, however, are not included in this analysis but do add to the assessment of the complete costs of coal.

To validate the findings, a life cycle assessment of coal-derived electricity was also performed using the Ecoinvent database in SimaPro v 7.1.²⁵ Health-related impact pathways were monetized using the value of disability-adjusted life-years from ExternE,²⁶ and the social costs of carbon.²⁰ Due to data limitations, this method could only be used to validate damages due to a subset of endpoints.

Box 2.

Summary Stats

1. Coal accounted for 25% of global energy consumption in 2005, but generated 41% of the CO₂ emissions that year.
2. In the United States, coal produces just over 50% of the electricity, but generates over 80% of the CO₂ emissions from the utility sector.²
3. Coal burning produces one and a half times more CO₂ emissions than does burning oil and twice that from burning natural gas (to produce an equal amount of energy).
4. The energy penalty from CCS (25–40%) would increase the amount of coal mined, transported, processed, and combusted, and the waste generated.⁴
5. Today, 70% of rail traffic in the United States is dedicated to shipping coal.²⁰ Land and transport would be further stressed with greater dependence on coal.

Life cycle impacts of coal

The health and environmental hazards associated with coal stem from extraction, processing, transportation and combustion of coal; the aerosolized,

solid, and liquid waste stream associated with mining, processing, and combustion; and the health, environmental, and economic impacts of climate change (Table 1).

Underground mining and occupational health

The U.S. Mine Safety and Health Administration (MSHA) and the National Institute for Occupational Safety and Health (NIOSH) track occupational injuries and disabilities, chronic illnesses, and mortality in miners in the United States. From 1973 to 2006 the incidence rate of all nonfatal injuries decreased from 1973 to 1987, then increased dramatically in 1988, then decreased from 1988 to 2006.²⁷ Major accidents still occur. In January 2006, 17 miners died in Appalachian coal mines, including 12 at the Sago mine in West Virginia, and 29 miners died at the Upper Big Branch Mine in West VA on April 5, 2010. Since 1900 over 100,000 have been killed in coal mining accidents in the United States.¹⁴

In China, underground mining accidents cause 3,800–6,000 deaths annually,²⁸ though the number of mining-related deaths has decreased by half over the past decade. In 2009, 2,631 coal miners were killed by gas leaks, explosions, or flooded tunnels, according to the Chinese State Administration of Work Safety.²⁹

Black lung disease (or pneumoconiosis), leading to chronic obstructive pulmonary disease, is the primary illness in underground coal miners. In the 1990s, over 10,000 former U.S. miners died from coal workers' pneumoconiosis and the prevalence has more than doubled since 1995.³⁰ Since 1900 coal workers' pneumoconiosis has killed over 200,000 in the United States.¹⁴ These deaths and illnesses are reflected in wages and workers' comp, costs considered internal to the coal industry, but long-term support often depends on state and federal funds.

Again, the use of "coking" coal used in industry is also omitted from this analysis: a study performed in Pittsburgh demonstrated that rates of lung cancer for those working on a coke oven went up two and one-half times, and those working on the top level had the highest (10-fold) risk.³¹

Mountaintop removal

MTR is widespread in eastern Kentucky, West Virginia, and southwestern Virginia. To expose coal seams, mining companies remove forests and fragment rock with explosives. The rubble or "spoil"

then sits precariously along edges and is dumped in the valleys below. MTR has been completed on approximately 500 sites in Kentucky, Virginia, West Virginia, and Tennessee,³² completely altering some 1.4 million acres, burying 2,000 miles of streams.³³ In Kentucky, alone, there are 293 MTR sites, over 1,400 miles of streams damaged or destroyed, and 2,500 miles of streams polluted.^{34–36} Valley fill and other surface mining practices associated with MTR bury headwater streams and contaminate surface and groundwater with carcinogens and heavy metals¹⁶ and are associated with reports of cancer clusters,³⁷ a finding that requires further study.

The deforestation and landscape changes associated with MTR have impacts on carbon storage and water cycles. Life cycle GHG emissions from coal increase by up to 17% when those from deforestation and land transformation by MTR are included.³⁸ Fox and Campbell estimated the resulting emissions of GHGs due to land use changes in the Southern Appalachian Forest, which encompasses areas of southern West Virginia, eastern Kentucky, southwestern Virginia, and portions of eastern Tennessee, from a baseline of existing forestland.³⁸ They estimated that each year, between 6 and 6.9 million tons of CO₂e are emitted due to removal of forest plants and decomposition of forest litter, and possibly significantly more from the mining “spoil” and lost soil carbon.

The fate of soil carbon and the fate of mining spoil, which contains high levels of coal fragments, termed “geogenic organic carbon,” are extremely uncertain and the results depend on mining practices at particular sites; but they may represent significant emissions. The Fox and Campbell³⁸ analysis determined that the worst-case scenario is that all soil carbon is lost and that all carbon in mining spoil is emitted—representing emissions of up to 2.6 million tons CO₂e from soil and 27.5 million tons CO₂e from mining spoil. In this analysis, the 6 million tons CO₂e from forest plants and forest litter represents our low and best estimates for all coal use, and 37 million tons CO₂e (the sum of the high bound of forest plants and litter, geogenic organic carbon, and the forest soil emissions) represents our high, upper bound estimate of emissions for all coal use. In the years Fox and Campbell studied, 90.5% of coal was used for electricity, so we attribute 90.5% of these emissions to coal-derived power.² To mon-

etize and bound our estimate for damages due to emissions from land disturbance, our point estimate for the cost was calculated using a social cost of carbon of \$30/ton CO₂e and our point estimate for emissions; the high-end estimate was calculated using the high-end estimate of emissions and a social cost of carbon of \$100/ton CO₂e; and the low estimate was calculated using the point estimate for emissions and the \$10/ton low estimate for the social cost of carbon.²⁰ Our best estimate is therefore \$162.9 million, with a range from \$54.3 million and \$3.35 billion, or 0.008¢/kWh, ranging from 0.003 ¢/kWh to 0.166 ¢/kWh.

The physical vulnerabilities for communities near MTR sites include mudslides and dislodged boulders and trees, and flash floods, especially following heavy rain events. With climate change, heavy rainfall events (2, 4, and 6 inches/day) have increased in the continental United States since 1970, 14%, 20%, and 27% respectively.^{39,40}

Blasting to clear mountain ridges adds an additional assault to surrounding communities.¹⁶ The blasts can damage houses, other buildings, and infrastructure, and there are numerous anecdotal reports that the explosions and vibrations are taking a toll on the mental health of those living nearby.

Additional impacts include losses in property values, timber resources, crops (due to water contamination), plus harm to tourism, corrosion of buildings and monuments, dust from mines and explosions, ammonia releases (with formation of ammonium nitrate), and releases of methane.⁴¹

Methane

In addition to being a heat-trapping gas of high potency, methane adds to the risk of explosions, and fires at mines.^{20,42} As of 2005, global atmospheric methane levels were approximately 1,790 parts per billion (ppb), which is an 27 ppb increase over 1998.⁴³ Methane is emitted during coal mining and it is 25 times more potent than CO₂ during a 100-year timeframe (this is the 100-year global warming potential, a common metric in climate science and policy used to normalize different GHGs to carbon equivalence). When methane decays, it can yield CO₂, an effect that is not fully assessed in this equivalency value.⁴³

According to the EIA,² 71,100,000 tons CO₂e of methane from coal were emitted in 2007 but

Table 1. The life cycle impact of the U.S. coal industry

	Economic	Human health	Environment	Other
Underground coal mining	1. Federal and state subsidies of coal industry	1. Increased mortality and morbidity in coal communities due to mining pollution 2. Threats remaining from abandoned mine lands	1. Methane emissions from coal leading to climate change 2. Remaining damage from abandoned mine lands	
MTR mining	1. Tourism loss 2. Significantly lower property values 3. Cost to taxpayers of environmental mitigation and monitoring (both mining and disposal stages) 4. Population declines	1. Contaminated streams 2. Direct trauma in surrounding communities 3. Additional mortality and morbidity in coal communities due to increased levels of air particulates associated with MTR mining (vs. underground mining) 4. Higher stress levels	1. Loss of biodiversity 2. Sludge and slurry ponds 3. Greater levels of air particulates 4. Loss and contamination of streams	
Coal mining	1. Opportunity costs of bypassing other types of economic development (especially for MTR mining) 2. Federal and state subsidies of coal industry 3. Economic boom and bust cycle in coal mining communities 4. Cost of coal industry litigation	1. Workplace fatalities and injuries of coal miners 2. Morbidity and mortality of mine workers resulting from air pollution (e.g., black lung, silicosis) 3. Increased mortality and morbidity in coal communities due to mining pollution 4. Increased morbidity and mortality due to increased air particulates in communities proximate to MTR mining	1. Destruction of local habitat and biodiversity to develop mine site 2. Methane emissions from coal leading to climate change 3. Loss of habitat and streams from valley fill (MTR) 4. Acid mine drainage	1. Infrastructure damage due to mudslides following MTR 2. Damage to surrounding infrastructure from subsidence 3. Damages to buildings and other infrastructure due to mine blasting 4. Loss of recreation availability in coal mining communities

Continued

Table 1. Continued

	Economic	Human health	Environment	Other
	5. Damage to farmland and crops resulting from coal mining pollution	5. Hospitalization costs resulting from increased morbidity in coal communities	5. Incomplete reclamation following mine use	5. Population losses in abandoned coal-mining communities
		6. Local health impacts of heavy metals in coal slurry	6. Water pollution from runoff and waste spills	
	6. Loss of income from small scale forest gathering and farming (e.g., wild ginseng, mushrooms) due to habitat loss	7. Health impacts resulting from coal slurry spills and water contamination	7. Remaining damage from abandoned mine lands	
	7. Loss of tourism income	8. Threats remaining from abandoned mine lands; direct trauma from loose boulders and felled trees	8. Air pollution due to increased particulates from MTR mining	
	8. Lost land required for waste disposal	9. Mental health impacts		
	9. Lower property values for homeowners	10. Dental health impacts reported, possibly from heavy metals		
	10. Decrease in mining jobs in MTR mining areas	11. Fungal growth after flooding		
Coal transportation	1. Wear and tear on aging railroads and tracks	1. Death and injuries from accidents during transport	1. GHG emissions from transport vehicles	1. Damage to rail system from coal transportation
		2. Impacts from emissions during transport	2. Damage to vegetation resulting from air pollution	2. Damage to roadways due to coal trucks
Coal combustion	1. Federal and state subsidies for the coal industry	1. Increased mortality and morbidity due to combustion pollution	1. Climate change due to CO ₂ and NO _x derived N ₂ O emissions	1. Corrosion of buildings and monuments from acid rain
	2. Damage to farmland and crops resulting from coal combustion pollution	2. Hospitalization costs resulting from increased morbidity in coal communities	2. Environmental contamination as a result of heavy metal pollution (mercury, selenium, arsenic)	2. Visibility impairment from NO _x emissions

Continued

Table 1. Continued

	Economic	Human health	Environment	Other
		3. Higher frequency of sudden infant death syndrome in areas with high quantities of particulate pollution	3. Impacts of acid rain derived from nitrogen oxides and SO ₂	
		4. See Levy <i>et al.</i> ²¹	4. Environmental impacts of ozone and particulate emissions	
			5. Soil contamination from acid rain	
			6. Destruction of marine life from mercury pollution and acid rain	
			7. Freshwater use in coal powered plants	
Waste disposal		1. Health impacts of heavy metals and other contaminants in coal ash and other waste	1. Impacts on surrounding ecosystems from coal ash and other waste	
		2. Health impacts, trauma and loss of property following coal ash spills	2. Water pollution from runoff and fly ash spills	
Electricity transmission	1. Loss of energy in the combustion and transmission phases		1. Disturbance of ecosystems by utility towers and rights of way	1. Vulnerability of electrical grid to climate change associated disasters

only 92.7% of this coal is going toward electricity. This results in estimated damages of \$2.05 billion, or 0.08¢/kWh, with low and high estimates of \$684 million and \$6.84 billion, or 0.034¢/kWh, and 0.34¢/kWh, using the low and high estimates for the social cost of carbon.²⁰ Life cycle assessment results, based on 2004 data and emissions from a subset of power plants, indicated 0.037 kg of CO₂e of methane emitted per kWh of electricity produced. With the best estimate for the social cost of carbon, this leads to an estimated cost of \$2.2 billion, or 0.11¢/kWh. The differences are due to differences in data, and

data from a different years. (See Fig. 1 for summary of external costs per kWh.)

Impoundments

Impoundments are found all along the periphery and at multiple elevations in the areas of MTR sites; adjacent to coal processing plants; and as coal combustion waste (“fly ash”) ponds adjacent to coal-fired power plants.⁴⁷ Their volume and composition have not been calculated.⁴⁸ For Kentucky, the number of known waste and slurry ponds alongside MTR sites and processing plants is 115.⁴⁹ These

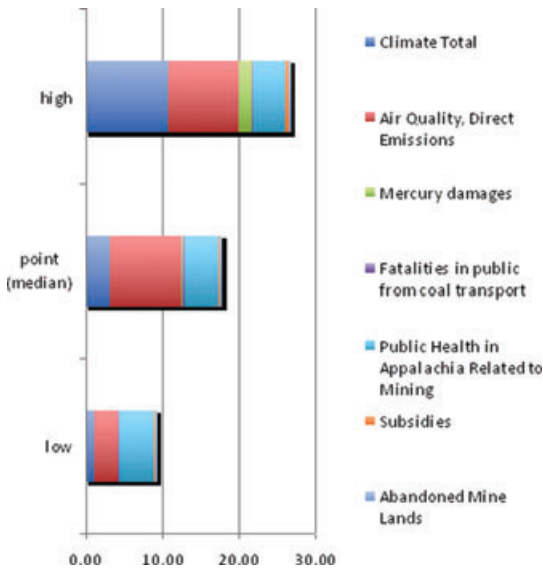


Figure 1. This graph shows the best estimates of the externalities due to coal, along with low and high estimates, normalized to ¢ per kWh of electricity produced. (In color in *Annals* online.)

sludge, slurry and coal combustion waste (CCW) impoundments are considered by the EPA to be significant contributors to water contamination in the United States. This is especially true for impoundments situated atop previously mined and potentially unstable sites. Land above tunnels dug for long-haul and underground mining are at risk of caving. In the face of heavier precipitation events, unlined containment dams, or those lined with dried slurry are vulnerable to breaching and collapse (Fig. 2).

Processing plants

After coal is mined, it is washed in a mixture of chemicals to reduce impurities that include clay, non-carbonaceous rock, and heavy metals to prepare for use in combustion.⁵⁰ Coal slurry is the by-product of these coal refining plants. In West Virginia, there are currently over 110 billion gallons of coal slurry permitted for 126 impoundments.^{49,51} Between 1972 and 2008, there were 53 publicized coal slurry spills in the Appalachian region, one of the largest of which was a 309 million gallon spill that occurred in Martin County, KY in 2000.⁴⁸ Of the known chemicals used and generated in processing coal, 19 are known cancer-causing agents, 24 are linked to lung and heart damage, and several remain untested as to their health effects.^{52,53}

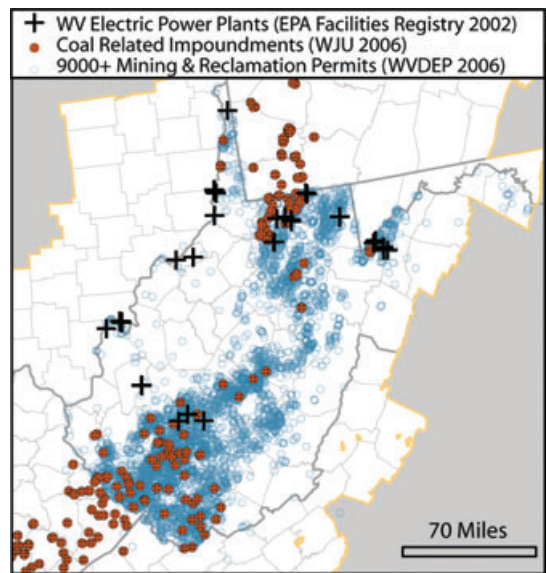


Figure 2. Electric power plants, impoundments (sludge and slurry ponds, CCW, or “fly ash”), and sites slated for reclamation in West Virginia.^{44–46} (In color in *Annals* online.) Source: Hope Childers, Wheeling Jesuit University.

Coal combustion waste or fly ash

CCW or fly ash—composed of products of combustion and other solid waste—contains toxic chemicals and heavy metals; pollutants known to cause cancer, birth defects, reproductive disorders, neurological damage, learning disabilities, kidney disease, and diabetes.^{47,54} A vast majority of the over 1,300 CCW impoundment ponds in the United States are poorly constructed, increasing the risk that waste may leach into groundwater supplies or nearby bodies of water.⁵⁵ Under the conditions present in fly ash ponds, contaminants, particularly arsenic, antimony, and selenium (all of which can have serious human health impacts), may readily leach or migrate into the water supplied for household and agricultural use.⁵⁶

According to the EPA, annual production of CCW increased 30% per year between 2000 and 2004, to 130 million tons, and is projected to increase to over 170 million tons by 2015.⁵⁷ Based on a series of state estimates, approximately 20% of the total is injected into abandoned coal mines.⁵⁸

In Kentucky, alone, there are 44 fly ash ponds adjacent to the 22 coal-fired plants. Seven of these ash ponds have been characterized as “high hazard”

by the EPA, meaning that if one of these impoundments spilled, it would likely cause significant property damage, injuries, illness, and deaths. Up to 1 in 50 residents in Kentucky, including 1 in 100 children, living near one of the fly ash ponds are at risk of developing cancer as a result of water- and air-borne exposure to waste.⁴⁷

Box 3.

Tennessee Valley Authority Fly Ash Pond Spill

On December 2, 2008 an 84-acre CCW containment area spilled when the dike ruptured at the Tennessee Valley Authority Kingston Fossil Plant CCW impoundment, following heavy rains. Over one billion gallons of fly ash slurry spilled across 300 acres.

Local water contamination

Over the life cycle of coal, chemicals are emitted directly and indirectly into water supplies from mining, processing, and power plant operations. Chemicals in the waste stream include ammonia, sulfur, sulfate, nitrates, nitric acid, tars, oils, fluorides, chlorides, and other acids and metals, including sodium, iron, cyanide, plus additional unlisted chemicals.^{16,50}

Spath and colleagues⁵⁰ found that these emissions are small in comparison to the air emissions. However, a more recent study performed by Koornneef and colleagues⁵⁹ using up-to-date data on emissions and impacts, found that emissions and seepage of toxins and heavy metals into fresh and marine water were significant. Elevated levels of arsenic in drinking water have been found in coal mining areas, along with ground water contamination consistent with coal mining activity in areas near coal mining facilities.^{16,17,60,61} In one study of drinking water in four counties in West Virginia, heavy metal concentrations (thallium, selenium, cadmium, beryllium, barium, antimony, lead, and arsenic) exceeded drinking water standards in one-fourth of the households.⁴⁸ This mounting evidence indicates that more complete coverage of water sampling is needed throughout coal-field regions.

Carcinogen emissions

Data on emissions of carcinogens due to coal mining and combustion are available in the Ecoin-

vent database.²⁵ The eco-indicator impact assessment method was used to estimate health damages in disability-adjusted life years due to these emissions,²⁵ and were valued using the VSL-year.²⁶ This amounted to \$11 billion per year, or 0.6 ¢/kWh, though these may be significant underestimates of the cancer burden associated with coal.

Of the emissions of carcinogens in the life cycle inventory (inventory of all environmental flows) for coal-derived power, 94% were emitted to water, 6% to air, and 0.03% were to soil, mainly consisting of arsenic and cadmium (note: these do not sum to 100% due to rounding).²⁵ This number is not included in our total cost accounting to avoid double counting since these emissions may be responsible for health effects observed in mining communities.

Mining and community health

A suite of studies of county-level mortality rates from 1979–2004 by Hendryx found that all-cause mortality rates,⁶² lung cancer mortality rates,⁶⁰ and mortality from heart, respiratory, and kidney disease¹⁷ were highest in heavy coal mining areas of Appalachia, less so in light coal mining areas, lesser still in noncoal mining areas in Appalachia, and lowest in noncoal mining areas outside of Appalachia. Another study performed by Hendryx and Ahern¹⁸ found that self-reports revealed elevated rates of lung, cardiovascular and kidney diseases, and diabetes and hypertension in coal-mining areas. Yet, another study found that for pregnant women, residing in coal mining areas of West Virginia posed an independent risk for low birth weight (LBW) infants, raising the odds of an LBW's infant by 16% relative to women residing in counties without coal mining.⁶³ LBW and preterm births are elevated,⁶⁴ and children born with extreme LBW fare worse than do children with normal birth weights in almost all neurological assessments;⁶⁵ as adults, they have more chronic diseases, including hypertension and diabetes mellitus.⁶⁶ Poor birth outcomes are especially elevated in areas with MTR mining as compared with areas with other forms of mining.⁶⁷ MTR mining has increased in the areas studied, and is occurring close to population centers.⁶²

The estimated excess mortality found in coal mining areas is translated into monetary costs using the VSL approach. For the years 1997–2005, excess age-adjusted mortality rates in coal mining areas of Appalachia compared to national rates

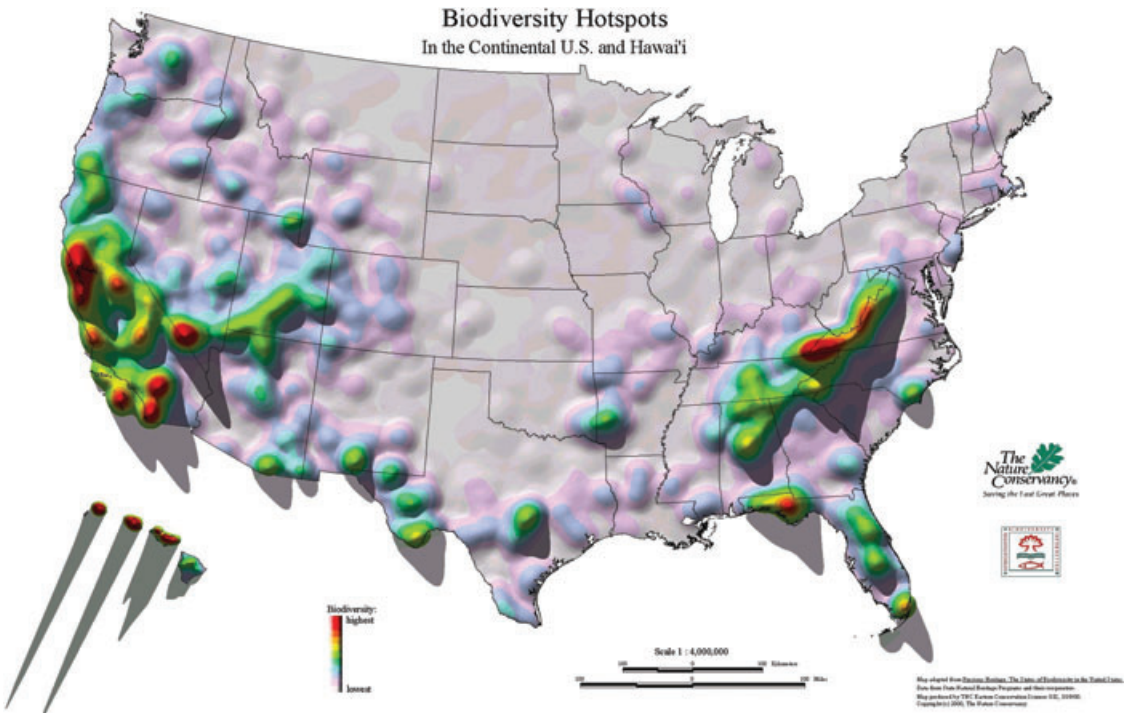


Figure 3. Areas of highest biological diversity in the continental United States. Source: The Nature Conservancy, Arlington, VA. (In color in *Annals* online.)

outside Appalachia translates to 10,923 excess deaths every year, with 2,347 excess deaths every year after, adjusting for other socio-economic factors, including smoking rates, obesity, poverty, and access to health care. These socio-economic factors were statistically significantly worse in coal-mining areas.^{18,62,68}

Using the VSL of \$7.5 million,²⁰ the unadjusted mortality rate, and the estimate that 91% of coal during these years was used for electricity,² this translates to a total cost of \$74.6 billion, or 4.36¢/kWh. In contrast, the authors calculated the direct (monetary value of mining industry jobs, including employees and proprietors), indirect (suppliers and others connected to the coal industry), and induced (ripple or multiplier effects throughout the economies) economic benefits of coal mining to Appalachia, and estimated the benefits to be \$8.08 billion in 2005 US\$.

Ecological impacts

Appalachia is a biologically and geologically rich region, known for its variety and striking beauty. There is loss and degradation of habitat from MTR;

impacts on plants and wildlife (species losses and species impacted) from land and water contamination, and acid rain deposition and altered stream conductivity; and the contributions of deforestation and soil disruption to climate change.^{16,20}

Globally, the rich biodiversity of Appalachian headwater streams is second only to the tropics.⁶⁹ For example, the southern Appalachian mountains harbor the greatest diversity of salamanders globally, with 18% of the known species world-wide (Fig. 3).⁶⁹

Imperiled aquatic ecosystems

Existence of viable aquatic communities in valley fill permit sites was first elucidated in court testimony leading to the “Haden decision.”⁷⁰ An interagency study of 30 streams in MTR mining-permit areas focused on the upper, unmapped reaches of headwater streams in West Virginia and Kentucky.⁷¹ In performing this study, the researchers identified 71 genera of aquatic insects belonging to 41 families within eight insect orders. The most widely distributed taxa in 175 samples were found in abundance in 30 streams in five areas slated to undergo MTR.

Electrical conductivity (a measure of the concentration of ions) is used as one indicator of stream health.⁷² The EPA recommends that stream conductivity not exceed 500 microsiemens per cm ($\mu\text{S}/\text{cm}$). In areas with the most intense mining, in which 92% of the watershed had been mined, a recent study revealed levels of 1,100 $\mu\text{S}/\text{cm}$.⁷²

Meanwhile, even levels below 500 $\mu\text{S}/\text{cm}$ were shown to significantly affect the abundance and composition of macroinvertebrates, such as mayflies and caddis flies.⁷³ “Sharp declines” were found in some stream invertebrates where only 1% of the watershed had been mined.^{74,75}

Semivoltine aquatic insects (e.g., many stoneflies and dragonflies)—those that require multiple years in the larval stage of development—were encountered in watersheds as small as 10–50 acres. While many of these streams become dry during the late summer months, they continue to harbor permanent resident taxonomic groups capable of withstanding summer dry conditions. Salamanders, the top predatory vertebrates in these fishless headwater streams, depend on permanent streams for their existence.

Mussels are a sensitive indicator species of stream health. Waste from surface mines in Virginia and Tennessee running off into the Clinch and Powell Rivers are overwhelming and killing these filter feeders, and the populations of mussels in these rivers has declined dramatically. Decreases in such filter feeders also affect the quality of drinking water downstream.⁷⁶

In addition, stream dwelling larval stages of aquatic insects are impossible to identify to the species level without trapping adults or rearing larvae to adults.⁷⁷ However, no studies of adult stages are conducted for mining-permit applications.

The view that—because there are so many small streams and brooks in the Appalachians—destroying a portion represents a minor threat to biodiversity is contrary to the science. As the planet’s second-oldest mountain range, geologically recent processes in Appalachia in the Pleistocene epoch (from 2.5 million to 12,000 years ago) have created conditions for diversification, resulting in one of the U.S. biodiversity “hotspots” (Fig. 3).

Thus, burying an entire 2,000 hectare watershed, including the mainstream and tributaries, is likely to eliminate species of multiple taxa found only in Appalachia.

Researchers have concluded that many unknown species of aquatic insects have likely been buried under valley fills and affected by chemically contaminated waterways. Today’s Appalachian coal mining is undeniably resulting in loss of aquatic species, many of which will never be known. Much more study is indicated to appreciate the full spectrum of the ecological effects of MTR mining.⁷⁸

Transport

There are direct hazards from transport of coal. People in mining communities report that road hazards and dust levels are intense. In many cases dust is so thick that it coats the skin, and the walls and furniture in homes.⁴¹ This dust presents an additional burden in terms of respiratory and cardiovascular disease, some of which may have been captured by Hendryx and colleagues.^{17–19,60,62,67,68,79}

With 70% of U.S. rail traffic devoted to transporting coal, there are strains on the railroad cars and lines, and (lost) opportunity costs, given the great need for public transport throughout the nation.²⁰

The NRC report²⁰ estimated the number of railroad fatalities by multiplying the proportion of revenue-ton miles (the movement of one ton of revenue-generating commodity over one mile) of commercial freight activity on domestic railroads accounted for by coal, by the number of public fatalities on freight railroads (in 2007); then multiplied by the proportion of transported coal used for electricity generation. The number of coal-related fatalities was multiplied by the VSL to estimate the total costs of fatal accidents in coal transportation. A total of 246 people were killed in rail accidents during coal transportation; 241 of these were members of the public and five of these were occupational fatalities. The deaths to the public add an additional cost of \$1.8 billion, or 0.09¢/kWh.

Social and employment impacts

In Appalachia, as levels of mining increase, so do poverty rates and unemployment rates, while educational attainment rates and household income levels decline.¹⁹

While coal production has been steadily increasing (from 1973 to 2006), the number of employees at the mines increased dramatically from 1973 to 1979, then decreased to levels below 1973 employment levels.²⁷ Between 1985 and 2005 employment in the Appalachian coal mining industry declined by 56% due to increases in mechanization for MTR and

other surface mining.^{19,27} There are 6,300 MTR and surface mining jobs in West Virginia, representing 0.7–0.8% of the state labor force.² Coal companies are also employing more people through temporary mining agencies and populations are shifting: between 1995 and 2000 coal-mining West Virginian counties experienced a net loss of 639 people to migration compared with a net migration gain of 422 people in nonmining counties.^{19,80}

Combustion

The next stage in the life cycle of coal is combustion to generate energy. Here we focus on coal-fired electricity-generating plants. The by-products of coal combustion include CO₂, methane, particulates and oxides of nitrogen, oxides of sulfur, mercury, and a wide range of carcinogenic chemicals and heavy metals.²⁰

Long-range air pollutants and air quality. Data from the U.S. EPA's Emissions & Generation Resource Integrated Database (eGRID)⁸¹ and National Emissions Inventory (NEI)⁸² demonstrates that coal power is responsible for much of the U.S. power generation-related emissions of PM_{2.5} (51%), NO_x (35%), and SO₂ (85%). Along with primary emissions of the particulates, SO₂ and NO_x contribute to increases in airborne particle concentrations through secondary transformation processes.^{20,21,83}

Studies in New England⁸⁴ find that, although populations within a 30-mile radius of coal-fired power plants make up a small contribution to aggregate respiratory illness, on a per capita basis, the impacts on those nearby populations are two to five times greater than those living at a distance. Data in Kentucky suggest similar zones of high impact.

The direct health impacts of SO₂ include respiratory illnesses—wheezing and exacerbation of asthma, shortness of breath, nasal congestion, and pulmonary inflammation—plus heart arrhythmias, LBW, and increased risk of infant death.

The nitrogen-containing emissions (from burning all fossil fuels and from agriculture) cause damages through several pathways. When combined with volatile organic compounds, they can form not only particulates but also ground-level ozone (photochemical smog). Ozone itself is corrosive to the lining of the lungs, and also acts as a local heat-trapping gas.

Epidemiology of air pollution. Estimates of non-fatal health endpoints from coal-related pollutants vary, but are substantial—including 2,800 from lung cancer, 38,200 nonfatal heart attacks and tens of thousands of emergency room visits, hospitalizations, and lost work days.⁸⁵ A review⁸³ of the epidemiology of airborne particles documented that exposure to PM_{2.5} is linked with all-cause premature mortality, cardiovascular and cardiopulmonary mortality, as well as respiratory illnesses, hospitalizations, respiratory and lung function symptoms, and school absences. Those exposed to a higher concentration of PM_{2.5} were at higher risk.⁸⁶ Particulates are a cause of lung and heart disease, and premature death,⁸³ and increase hospitalization costs. Diabetes mellitus enhances the health impacts of particulates⁸⁷ and has been implicated in sudden infant death syndrome.⁸⁸ Pollution from two older coal-fired power plants in the U.S. Northeast was linked to approximately 70 deaths, tens of thousands of asthma attacks, and hundreds of thousands of episodes of upper respiratory illnesses annually.⁸⁹

A reanalysis of a large U.S. cohort study on the health effects of air pollution, the Harvard Six Cities Study, by Schwartz *et al.*⁹⁰ used year-to-year changes in PM_{2.5} concentrations instead of assigning each city a constant PM_{2.5} concentration. To construct one composite estimate for mortality risk from PM_{2.5}, the reanalysis also allowed for yearly lags in mortality effects from exposure to PM_{2.5}, and revealed that the relative risk of mortality increases by 1.1 per 10 µg/m³ increase in PM_{2.5} the year of death, but just 1.025 per 10 µg/m³ increase in PM_{2.5} the year before death. This indicates that most of the increase in risk of mortality from PM_{2.5} exposure occurs in the same year as the exposure. The reanalysis also found little evidence for a threshold, meaning that there may be no “safe” levels of PM_{2.5} and that all levels of PM_{2.5} pose a risk to human health.⁹¹

Thus, prevention strategies should be focused on continuous reduction of PM_{2.5} rather than on peak days, and that air quality improvements will have effect almost immediately upon implementation. The U.S. EPA annual particulate concentration standard is set at 15.0 µg/m³, arguing that there is no evidence for harm below this level.⁹² The results of the Schwartz *et al.*⁹⁰ study directly contradict this line of reasoning.

Risk assessment. The risk assessment performed by the NRC,²⁰ found aggregate damages of \$65 billion, including damages to public health, property, crops, forests, foregone recreation, and visibility due to emissions from coal-fired power plants of PM_{2.5}, PM₁₀, SO₂, NO_x, volatile organic compounds, and ozone. The public health damages included mortality cases, bronchitis cases, asthma cases, hospital admissions related to respiratory, cardiac, asthma, coronary obstructive pulmonary disease, and ischemic heart disease problems, and emergency room visits related to asthma. On a plant-by-plant basis after being normalized to electricity produced by each plant, this was 3.2 ¢/kWh. Plant-by-plant estimates of the damages ranged from 1.9 ¢/kWh to 12 ¢/kWh. Plant-to-plant variation was largely due to controls on the plant, characteristics of the coal, and the population downwind of the plant. Emissions of SO₂ were the most damaging of the pollutants affecting air quality, and 99% of this was due to SO₂ in the particle form.²⁰ The NRC study found that over 90% of the damages due to air quality are from PM_{2.5}-related mortality, which implies that these damages included approximately 8,158 excess mortality cases.²⁰ For the state of Kentucky alone, for each ton of SO₂ removed from the stack, the NRC (2009)²⁰ calculated a public health savings of \$5,800. Removing the close to 500,000 tons emitted in Kentucky would save over \$2.85 billion annually. The life cycle analysis found that damages from air quality public health impacts, monetized using methods from Externe²⁶ are approximately \$70.5 billion, which is roughly in line with this number.

The NRC's estimate is likely an underestimate, since the NRC used the concentration-response curve from Pope and Dockery,⁸³ which provides a low estimate for increases in mortality risk with increases in PM_{2.5} exposure and is an outlier when compared to other studies examining the PM_{2.5}-mortality relationship.^{6,87} Had they used the result of the more recent study by Schwartz *et al.*,⁹⁰ which was used in a similar study by Levy *et al.*,²¹ or the number from Dockery *et al.*,⁹³ the value they calculated would have been approximately three times higher,²⁰ therefore implying 24,475 excess deaths in 2005, with a cost of \$187.5 billion, or 9.3¢/kWh. As the Schwartz *et al.* study is more recent, uses elaborate statistical techniques to derive the concentration-response function for PM_{2.5} and mortality, and is now widely accepted,^{21,94} we use it

here to derive our best and high estimate, and the Pope and Dockery,⁸³ estimate to derive our low. Our best and high estimates for the damages due to air quality detriment impacts are both \$187.5 billion, and our low is \$65 billion. On a per-kWh basis, this is an average cost of 9.3 ¢/kWh with a low estimate of 3.2 ¢/kWh.

Atmospheric nitrogen deposition. In addition to the impacts to air quality and public health, nitrogen causes ecological harm via eutrophication. Eutrophication, caused by excess nitrogen inputs to coastal river zones, is the greatest source of water quality alteration in the United States and atmospheric deposition is one of the dominant sources of nitrogen inputs.⁹⁵ In an analysis by Jaworski *et al.*,⁹⁵ prepared for the EPA, 10 benchmark watersheds in the U.S. Northeast that flowed into the Atlantic coastal zone with good historical data were analyzed in conjunction with emissions data and reconstructed historical emissions. They found that the contribution to riverine nitrogen from nitrogen deposited from the air ranged from 36% to 80%, with a mean of 64%.

The other primary sources of nitrogen are fertilizers from point (e.g., river) discharges and nonpoint (e.g., agricultural land) sources, and other point sources including sewage from cities and farm animals, especially concentrated animal feeding operations.⁹⁵ Anthropogenic contributions of nitrogen are equal to the natural sources, doubling this form of fertilization of soils and water bodies.⁹⁶

Harmful algal blooms and dead zones

Ocean and water changes are not usually associated with coal. But nitrogen deposition is a by-product of combustion and the EPA⁹⁷ has reached consensus on the link between aquatic eutrophication and harmful algal blooms (HABs), and concluded that nutrient over-fertilization is one of the reasons for their expansion in the United States and other nations. HABs are characterized by discolored water, dead and dying fish, and respiratory irritants in the air, and have impacts including illness and death, beach closures, and fish, bird, and mammal die-offs from exposure to toxins. Illnesses in humans include gastroenteritis, neurological deficits, respiratory illness, and diarrhetic, paralytic, and neurotoxic shellfish poisonings.

N₂O from land clearing is a heat-trapping gas^{38,42} and adds to the nitrogen deposited in soils and water

bodies. The nitrogen is also a contributor to fresh and sea water acidification.^{98–100} Other factors include the loss of wetlands that filter discharges.^{98–100}

The economic losses from HABs are estimated to be over \$82 million/year in the United States, based on the most prominent episodes.^{101,102} The full economic costs of HABs include public health impacts and health care costs, business interruptions of seafood and other allied industries (such as tourism and recreation, unemployment of fin- and shellfish fisherman and their families), and disruptions of international trade.^{98–100}

The overfertilization of coastal zones worldwide has also led to over 350 “dead zones” with hypoxia, anoxia, and death of living marine organisms. Commercial and recreational fisheries in the Gulf of Mexico generate \$2.8 billion annually¹⁰³ and losses from the heavily eutrophied Gulf of Mexico dead zone put the regional economy at risk.

Acid precipitation. In addition to the health impacts of SO₂, sulfates contribute to acid rain, decreased visibility, and have a greenhouse cooling influence.²⁰

The long-term Hubbard Brook Ecosystem Study¹⁰⁴ has demonstrated that acid rain (from sulfates and nitrates) has taken a toll on stream and lake life, and soils and forests in the United States, primarily in the Northeast. The leaching of calcium from soils is widespread and, unfortunately, the recovery time is much longer than the time it takes for calcium to become depleted under acidic conditions.¹⁰⁵

No monetized values of costs were found but a value for the benefits of improvements to the Adirondack State Park from acid rain legislation was produced by Resources for the Future, and found benefits ranging from \$336 million to \$1.1 billion per year.¹⁰⁶

Mercury. Coal combustion in the U.S. releases approximately 48 tons of the neurotoxin mercury each year.⁵⁴ The most toxic form of mercury is methylmercury, and the primary route of human exposure is through consumption of fin- and shellfish containing bioaccumulated methylmercury.¹⁰⁷ Methylmercury exposure, both dietary and *in utero* through maternal consumption, is associated with neurological effects in infants and children, including delayed achievement of developmental milestones and poor results on neurobehavioral

tests—attention, fine motor function, language, visual-spatial abilities, and memory. Seafood consumption has caused 7% of women of childbearing age to exceed the mercury reference dose set by the EPA, and 45 states have issued fish consumption advisories.¹⁰⁷ Emission controls specific to mercury are not available, though 74–95% of emitted mercury is captured by existing emissions control equipment. More advanced technologies are being developed and tested.¹⁰⁷

Direct costs of mercury emissions from coal-fired power plants causing mental retardation and lost productivity in the form of IQ detriments were estimated by Trasande *et al.*^{22,23} to be \$361.2 million and \$1.625 billion, respectively, or 0.02¢/kWh and 0.1¢/kWh, respectively. Low-end estimates for these values are \$43.7 million and \$125 million, or 0.003¢/kWh and 0.007¢/kWh; high-end estimates for these values are \$3.3 billion and \$8.1 billion, or 0.19¢/kWh and 0.48¢/kWh.

There are also epidemiological studies suggesting an association between methylmercury exposure and cardiovascular disease.¹⁰⁸ Rice *et al.*¹⁰⁹ monetized the benefits of a 10% reduction in mercury emissions for both neurological development and cardiovascular health, accounting for uncertainty that the relationship between cardiovascular disease and methylmercury exposure is indeed causal. Applying these results for the cardiovascular benefits of a reduction in methylmercury to the 41% of total U.S. mercury emissions from coal^{22,23} indicates costs of \$3.5 billion, with low and high estimates of \$0.2 billion and \$17.9 billion, or 0.2 ¢/kWh, with low and high estimates of 0.014 ¢/kWh and 1.05 ¢/kWh.

Coal's contributions to climate change

The Intergovernmental Panel on Climate Change (IPCC) reported that annual global GHG emissions have—between 1970 and 2004—increased 70% to 49.0 Gt CO₂-e/year.¹⁰⁹ The International Energy Agency's Reference Scenario estimates that worldwide CO₂ emissions will increase by 57% between 2005 and 2030, or 1.8% each year, to 41,905 Mt.¹ In the same time period, CO₂ emissions from coal-generated power are projected to increase 76.6% to 13,884 Mt.¹

In 2005, coal was responsible for 82% of the U.S.'s GHG emissions from power generation.¹¹⁰ In addition to direct stack emissions, there are methane

emissions from coal mines, on the order of 3% of the stack emissions.¹¹⁰ There are also additional GHG emissions from the other uses of coal, approximately 139 Mt CO₂.¹

Particulate matter (black carbon or soot) is also a heat-trapping agent, absorbing solar radiation, and, even at great distances, decreasing reflectivity (albedo) by settling in snow and ice.^{111–113} The contribution of particulates (from coal, diesel, and biomass burning) to climate change has, until recently, been underestimated. Though short-lived, the global warming potential per volume is 500 times that of CO₂.¹¹¹

Climate change

Since the 1950s, the world ocean has accumulated 22 times as much heat as has the atmosphere,¹¹⁴ and the pattern of warming is unmistakably attributable to the increase in GHGs.¹¹⁵ Via this ocean repository and melting ice, global warming is changing the climate: causing warming, altered weather patterns, and sea level rise. Climate may change gradually or nonlinearly (in quantum jumps). The release of methane from Arctic seas and the changes in Earth's ice cover (thus albedo), are two potential amplifying feedbacks that could accelerate the rate of Earth's warming.

Just as we have underestimated the rate at which the climate would change, we have underestimated the pace of health and environmental impacts. Already the increases in asthma, heat waves, clusters of illnesses after heavy rain events and intense storms, and in the distribution of infectious diseases are apparent.^{116,117} Moreover, the unfolding impacts of climate instability hold yet even more profound impacts for public health, as the changes threaten the natural life-supporting systems upon which we depend.

The EIA² estimated that 1.97 billion tons of CO₂ and 9.3 million tons CO₂e of N₂O were emitted directly from coal-fired power plants. Using the social cost of carbon, this resulted in a total cost of \$61.7 billion, or 3.06 ¢/kWh. Using the low and high estimates of the social cost of carbon results in cost of \$20.56 billion to \$205.6 billion, or 1.02 ¢/kWh to 10.2 ¢/kWh.

Black carbon emissions were also calculated using data from the EPA's eGRID database⁸¹ on electricity produced from lignite. The low, mean, and high energy density values for lignite⁵ was then used

to calculate the amount of lignite consumed. The Cooke *et al.*¹¹⁸ emissions factor was used to estimate black carbon emissions based on lignite use and the Hansen *et al.*¹¹¹ global temperature potential was used to convert these emissions to CO₂e. This resulted in an estimate of 1.5 million tons CO₂e being emitted in 2008, with a value of \$45.2 million, or 0.002¢/kWh. Using our low and high estimates for the social cost of carbon and the high and low values for the energy density of lignite produced values of \$12.3 million to \$161.4 million, or 0.0006 ¢/kWh to 0.008¢/kWh.

One measure of the costs of climate change is the rising costs of extreme weather events, though these are also a function of and real estate and insurance values. Overall, the costs of weather-related disasters rose 10-fold from the 1980s to the 1990s (from an average of \$4 bn/year to \$40 bn/year) and jumped again in the past decade, reaching \$225 bn in 2005.¹¹⁹ Worldwide, Munich Re—a company that insures insurers—reports that, in 2008, without Katrina-level disasters, weather-related “catastrophic losses” to the global economy were the third-highest in recorded history, topping \$200 billion, including \$45 billion in the United States.¹²⁰

The total costs of climate change damages from coal-derived power, including black carbon, CO₂ and N₂O emissions from combustion, land disturbance in MTR, and methane leakage from mines, is \$63.9 billion dollars, or 3.15 ¢/kWh, with low and high estimates of \$21.3 billion to \$215.9 billion, or 1.06 ¢/kWh to 10.71 ¢/kWh. A broad examination of the costs of climate change¹²¹ projects global economic losses to between 5 and 20% of global gross domestic product (\$1.75–\$7 trillion in 2005 US\$); the higher figure based on the potential collapse of ecosystems, such as coral reefs and widespread forest and crop losses. With coal contributing at least one-third of the heat-trapping chemicals, these projections offer a sobering perspective on the evolving costs of coal; costs that can be projected to rise (linearly or nonlinearly) over time.

Carbon capture and storage

Burning coal with CO₂ CCS in terrestrial, ocean, and deep ocean sediments are proposed methods of deriving “clean coal.” But—in addition to the control technique not altering the upstream life cycle costs—significant obstacles lie in the way, including the costs of construction of suitable plants

Table 2. MIT cost estimates for some representative CCS systems.⁵

		Subcritical PC		Supercritical PC		Ultra-supercritical PC		SC PC-Oxy	IGCC	
		No capture	Capture	No capture	Capture	No capture	Capture	Capture	No capture	Capture
CCS performance	Coal feed (kg/hr)	208,000	284,000	184,894	242,950	164,000	209,000	232,628	185,376	228,115
	CO ₂ emitted (kg/hr)	466,000	63,600	414,903	54,518	369,000	46,800	52,202	415,983	51,198
	CO ₂ captured at 90%, (kg/h)	0	573,000	0	490,662	0	422,000	46,981.7	0	460,782
	CO ₂ emitted (g/kWh)	931	127	830	109	738	94	104	832	102
CCS costs	\$/kWh	1,280	2,230	1,330	2,140	1,360	2,090	1,900	1,430	1,890
	Total \$, assuming 500 MW plant	\$640,000,000	\$1,115,000,000	\$665,000,000	\$1,070,000,000	\$680,000,000	\$1,045,000,000	\$950,000,000	\$715,000,000	\$945,000,000
	Inv. Charge ¢/kWh @ 15.1%	2.6	4.52	2.7	4.34	2.76	4.24	3.85	2.9	3.83
	Fuel ¢/kWh @ \$1.50/MMBtu	1.49	2.04	1.33	1.75	1.18	1.5	1.67	1.33	1.64
	O&M ¢/kWh	0.75	1.6	0.75	1.6	0.75	1.6	1.45	0.9	1.05
	COE ¢/kWh	4.84	8.16	4.78	7.69	4.69	7.34	8.98	5.13	6.52
	Cost of CO ₂ avoided vs. same technology w/o capture (\$/ton)		41.3		40.4		41.1	30.3		19.3
	Cost of CO ₂ avoided vs. supercritical technology w/o capture (\$/ton)		48.2		40.4		34.8	30.3		24
	Energy penalty		1,365,384,615		1,313,996,128		1,274,390,244		1,230,553,038	

and underground storage facilities, and the “energy penalty” requiring that coal consumption per unit of energy produced by the power plant increase by 25–40% depending on the technologies used.^{4,42}

Retrofitting old plants—the largest source of CO₂ in the United States—may exact an even larger energy penalty. The energy penalty means that more coal is needed to produce the same quantity of electricity, necessitating more mining, processing, and transporting of coal and resulting in a larger waste stream to produce the same amount of electricity. Coal-fired plants would still require locally polluting diesel trucks to deliver the coal, and generate CCW ponds that can contaminate ground water. Given current siting patterns, such impacts often fall disproportionately on economically disadvantaged communities. The energy penalty combined with other increased costs of operating a CCS plant would nearly double the cost of generating electricity from that plant, depending on the technology used (see Table 2).⁵

The U.S. Department of Energy estimates that an underground volume of 30,000 km² will be needed per year to reduce the CO₂ emissions from coal by 20% by 2050 (the total land mass of the continental U.S. (48 states) is 9,158,960 km²).¹²²

The safety and ensurability of scaling up the storage of the billion tons of CO₂ generated each year into the foreseeable future are unknown. Extrapolating from localized experiments, injecting fractions of the volumes that will have to be stored to make a significant difference in emissions, is fraught with numerous assumptions. Bringing CCS to scale raises additional risks, in terms of pressures underground. In addition to this, according to the U.S. Government Accountability Office (2008) there are regulatory, legal and liability uncertainties, and there is “significant cost of retrofitting existing plants that are single largest source of CO₂ emissions in the United States” (p. 7).¹²³

Health and environmental risks of CCS

The Special IPCC Report on Carbon Dioxide Capture and Storage⁴² lists the following concerns for CCS in underground terrestrial sites:

1. Storing compressed and liquefied CO₂ underground can acidify saline aquifers (akin to ocean acidification) and leach heavy metals, such as arsenic and lead, into ground water.⁴²
2. Acidification of ground water increases fluid-rock interactions that enhance calcite dissolution and solubility, and can lead to fractures in

limestone (CaCO_3) and subsequent releases of CO_2 in high concentrations.¹²⁴

3. Increased pressures may cause leaks and releases from previously drilled (often unmapped) pathways.
4. Increased pressures could destabilize underground faults and lead to earthquakes.
5. Large leaks and releases of concentrated CO_2 are toxic to plants and animals.⁴²
 - a. The 2006 Mammoth Mountain, CA release left dead stands of trees.¹²⁴
6. Microbial communities may be altered, with release of other gases.⁴²

The figures in Table 2 represent costs for new construction. Costs for retrofits (where CCS is installed on an active plant) and rebuilds (where CCS is installed on an active plant and the combustion technology is upgraded) are highly uncertain because they are extremely dependent on site conditions and precisely what technology the coal plant is upgraded to.⁵ It does appear that complete rebuilds are more economically attractive than retrofits, and that “carbon-capture ready” plants are not economically desirable to build.⁵

Subsidies

In Kentucky, coal brings in an estimated \$528 million in state revenues, but is responsible for \$643 million in state expenditures. The net impact, therefore, is a loss of \$115 million to the state of Kentucky.¹²⁶ These figures do not include costs of health care, lost productivity, water treatment for siltation and water infrastructure, limited development potential due to poor air quality, and social expenditures associated with declines in employment and related economic hardships of coal-field communities.¹²⁶

The U.S. Federal Government provides subsidies for electricity and mining activities, and these have been tallied by both the EIA and the Environmental Law Institute.^{2,127,128} The EIA estimate is \$3.17 billion of subsidies in 2007, or 0.16¢/kWh, and the Environmental Law Institute estimate is \$5.37 billion for 2007, or 0.27¢/kWh.

Abandoned mine lands

Abandoned mine lands (AML) are those lands and waters negatively impacted by surface coal mining and left inadequately reclaimed or abandoned prior to August 3, 1977.¹²⁹ There are over 1,700 old aban-

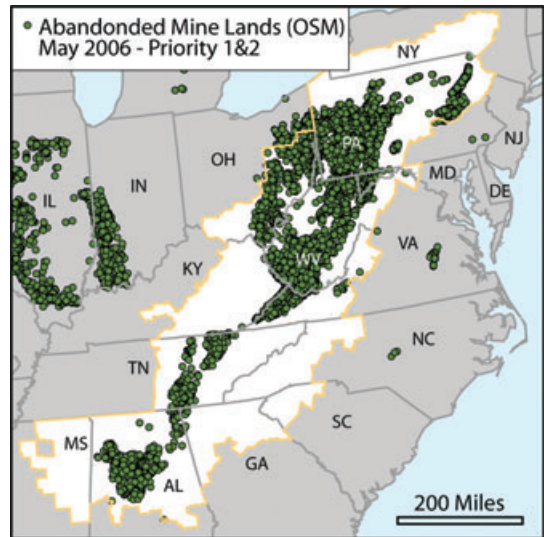


Figure 4. Current high-priority abandoned mine land reclamation sites from Alabama to Pennsylvania.¹²⁹ (In color in *Animals* online.) Source: Hope Childers, Wheeling Jesuit University.

doned mines in Pennsylvania, alone.¹⁴ In some—like that in Centralia, PA—fires burn for decades, emitting carbon monoxide, and other fumes. The ground above others can open, and several people die each year falling into them. Still others flood and lead to contaminated ground water. Previous coal mining communities lie in the shadow of these disturbed areas. Officials in Pennsylvania estimate that it will take \$15 billion over six decades to clean Pennsylvania’s abandoned mines.

Since the passage of the Surface Mining Control and Reclamation Act of 1977, active mining operations have been required to pay fees into the Abandoned Mine Reclamation Fund that are then used to finance reclamation of these AMLs.¹²⁹ Despite the more than \$7.4 billion that has been collected as of September 30, 2005, there is a growing backlog of unfunded projects.⁵¹ Data on the number and monetary value of unfunded AML projects remaining at the end of 2007 for the nation were collected directly from the Abandoned Mine Land Inventory System¹²⁹ and amounted to \$8.8 billion 2008 US\$, or 0.44¢/kWh (Fig. 4).

Results

The tabulation of the externalities in total and converted to 2008 US\$ is given in Table 3 and normalized to cents per kWh of coal-generated electricity

Table 3. The complete costs of coal as reviewed in this report in 2008 US\$.

	Monetized estimates from literature (2008 US\$)			Monetized life cycle assessment results (2008 US\$)	
	Low	Best	High	IPCC 2007, U.S.	U.S. Hard Coal
				Hard Coal	Eco-indicator
Land disturbance	\$54,311,510	\$162,934,529	\$3,349,209,766		
Methane emissions from mines	\$684,084,928	\$2,052,254,783	\$6,840,849,276	\$2,188,192, 405	
Carcinogens (mostly to water from waste)					\$11,775,544, 263
Public health burden of communities in Appalachia	\$74,612,823,575	\$74,612,823,575	\$74,612,823,575		
Fatalities in the public due to coal transport	\$1,807,500,000	\$1,807,500,000	\$1,807,500,000		
Emissions of air pollutants from combustion	\$65,094,911,734	\$187,473,345,794	\$187,473,345,794		\$71,011,655, 364
Lost productivity from mercury emissions	\$125,000,000	\$1,625,000,000	\$8,125,000,000		
Excess mental retardation cases from mercury emissions	\$43,750,000	\$361,250,000	\$3,250,000,000		
Excess cardiovascular disease from mercury emissions	\$246,000,000	\$3,536,250,000	\$17,937,500,000		
Climate damages from combustion emissions of CO ₂ and N ₂ O	\$20,559,709,242	\$61,679,127,726	\$205,597,092,419.52	\$70,442,466, 509	
Climate damages from combustion emissions of black carbon	\$12,346,127	\$45,186,823	\$161,381,512.28	\$3,739,876, 478	
Environmental Law Institute estimate 2007			\$5,373, 963,368		
EIA 2007	\$3,177,964,157	\$3,177, 964,157			
AMLs	\$8,775,282,692	\$8,775, 282,692	\$8,775, 282,692		
Climate total	\$21,310,451,806	\$63,939,503,861	\$215,948,532,974		
Total	\$175,193,683,964	\$345,308,920,080	\$523,303,948,403		

A 2010 Clean Air Task Force⁵⁶ (CATF) report, with Abt Associates consulting, lists 13,000 premature deaths due to air pollution from all electricity generation in 2010, a decrease in their estimates from previous years. They attribute the drop to 105 scrubbers installed since 2005, the year in which we based our calculations. We were pleased to see improvements reported in air quality and health outcomes. There is, however, considerable uncertainty regarding the actual numbers. Using the epidemiology from the “Six Cities Study” implies up to 34,000 premature deaths in 2010. Thus, our figures are mid-range while those of the CATF represent the most conservative of estimates.

in Table 4. Our best estimate for the externalities related to coal is \$345.3 billion (range: \$175.2 bn to \$523.3 bn). On a per-kWh basis this is 17.84¢/kWh, ranging from 9.42 ¢/kWh to 26.89 ¢/kWh.

Limitations of this analysis

While we have based this analysis on the best available data that are used by a wide range of organizations, this review is limited by the omission of

many environmental, community, mental health, and economic impacts that are not easily quantifiable. Another limitation is the placing of numbers on impacts that are difficult to quantify or monetize, including the VSL, a crude estimate of the benefits of reducing the number of deaths used by economists, and the social cost of carbon, based on the evolving impacts of climate change. We have included ranges, reflecting the numerous sets of data and studies in this field (all of which have their own

Table 4. Total costs of coal normalized to kWh of electricity produced.

	Monetized estimates from literature in ¢/kWh of electricity (2008 US\$)			Monetized life cycle assessment results in ¢/kWh of electricity (2008 US\$)	
	Low	Best	High	IPCC 2007, U.S. Hard Coal	U.S. Hard Coal Eco-indicator
Land disturbance	0.00	0.01	0.17		
Methane emissions from mines	0.03	0.08	0.34	0.11	
Carcinogens (mostly to water from waste)					0.60
Public health burden of communities in Appalachia	4.36	4.36	4.36		
Fatalities in the public due to coal transport	0.09	0.09	0.09		
Emissions of air pollutants from combustion	3.23	9.31	9.31		3.59
Lost productivity from mercury emissions	0.01	0.10	0.48		
Excess mental retardation cases from mercury emissions	0.00	0.02	0.19		
Excess cardiovascular disease from mercury emissions	0.01	0.21	1.05		
Climate damage from combustion emissions of CO ₂ and N ₂ O	1.02	3.06	10.20	3.56	
Climate damages from combustion emissions of black carbon	0.00	0.00	0.01	0.19	
Environmental Law Institute estimate 2007			0.27		
EIA 2007	0.16	0.16			
AMLs	0.44	0.44	0.44		
Climate total	1.06	3.15	10.7	3.75	1.54
Total	9.36	17.84	26.89		

uncertainties), varying assumptions in data sets and studies, and uncertainties about future impacts and the costs to society.

Some of the issues raised apply only to the region discussed. Decreased tourism in Appalachia, for example, affects regional economies; but may not affect the overall economy of the United States, as tourists may choose other destinations.

Studies in Australian coal mining communities illustrate the cycle of economic boom during construction and operation, the economic and worker decoupling from the fortunes of the mines; then the eventual closing.¹³⁰ Such communities experience high levels of depression and poverty, and increases in assaults (particularly sexual assaults), motor vehicle accidents, and crimes against

property, until the culture shifts to allow for development of secondary industries. Additional evidence documents that mining-dependent economies tend to be weak economies,¹³¹ and weak economic conditions in turn are powerful predictors of social and health disadvantages.^{130,132}

Some values are also difficult to interpret, given the multiple baselines against which they must be compared. In assessing the “marginal” costs of environmental damages, we have assumed the diverse, pristine, hardwood forest that still constitutes the majority of the beautiful rich and rolling hills that make up the Appalachian Mountain range.

Ecological and health economic analyses are also affected by the discount rate used in such evaluations. Discount rates are of great value in assessing the worth of commodities that deteriorate over time. But they are of questionable value in assessing ecological, life-supporting systems that have value if they are sustained. Ecological economists might consider employing a negative discount rate—or an accrual rate—in assessing the true impacts of environmental degradation and the value of sustainability.

Finally, the costs reported here do not include a wide range of opportunity costs, including lost opportunities to construct wind farms and solar power plants, begin manufacture of wind turbines and solar technologies, develop technologies for the smart grid and transmission, and for economic and business development unrelated to the energy sector.

Conclusions

The electricity derived from coal is an integral part of our daily lives. However, coal carries a heavy burden. The yearly and cumulative costs stemming from the aerosolized, solid, and water pollutants associated with the mining, processing, transport, and combustion of coal affect individuals, families, communities, ecological integrity, and the global climate. The economic implications go far beyond the prices we pay for electricity.

Our comprehensive review finds that the best estimate for the total economically quantifiable costs, based on a conservative weighting of many of the study findings, amount to some \$345.3 billion, adding close to 17.8¢/kWh of electricity generated from coal. The low estimate is \$175 billion, or over 9¢/kWh, while the true monetizable costs could be as much as the upper bounds of \$523.3 billion,

adding close to 26.89¢/kWh. These and the more difficult to quantify externalities are borne by the general public.

Still these figures do not represent the full societal and environmental burden of coal. In quantifying the damages, we have omitted the impacts of toxic chemicals and heavy metals on ecological systems and diverse plants and animals; some ill-health endpoints (morbidity) aside from mortality related to air pollutants released through coal combustion that are still not captured; the direct risks and hazards posed by sludge, slurry, and CCW impoundments; the full contributions of nitrogen deposition to eutrophication of fresh and coastal sea water; the prolonged impacts of acid rain and acid mine drainage; many of the long-term impacts on the physical and mental health of those living in coal-field regions and nearby MTR sites; some of the health impacts and climate forcing due to increased tropospheric ozone formation; and the full assessment of impacts due to an increasingly unstable climate.

The true ecological and health costs of coal are thus far greater than the numbers suggest. Accounting for the many external costs over the life cycle for coal-derived electricity conservatively doubles to triples the price of coal per kWh of electricity generated.

Our analysis also suggests that the proposed measure to address one of the emissions—CO₂, via CCS—is costly and carries numerous health and environmental risks, which would be multiplied if CCS were deployed on a wide scale. The combination of new technologies and the “energy penalty” will, conservatively, almost double the costs to operate the utility plants. In addition, questions about the reserves of economically recoverable coal in the United States carry implications for future investments into coal-related infrastructure.

Public policies, including the Clean Air Act and New Source Performance Review, are in place to help control these externalities; however, the actual impacts and damages remain substantial. These costs must be accounted for in formulating public policies and for guiding private sector practices, including project financing and insurance underwriting of coal-fired plants with and without CCS.

Recommendations

1. Comprehensive comparative analyses of life cycle costs of all electricity generation

technologies and practices are needed to guide the development of future energy policies.

2. Begin phasing out coal and phasing in cleanly powered smart grids, using place-appropriate alternative energy sources.
3. A healthy energy future can include electric vehicles, plugged into cleanly powered smart grids; and healthy cities initiatives, including green buildings, roof-top gardens, public transport, and smart growth.
4. Alternative industrial and farming policies are needed for coal-field regions, to support the manufacture and installation of solar, wind, small-scale hydro, and smart grid technologies. Rural electric co-ops can help in meeting consumer demands.
5. We must end MTR mining, reclaim all MTR sites and abandoned mine lands, and ensure that local water sources are safe for consumption.
6. Funds are needed for clean enterprises, reclamation, and water treatment.
7. Fund-generating methods include:
 - a. maintaining revenues from the workers' compensation coal tax;
 - b. increasing coal severance tax rates;
 - c. increasing fees on coal haul trucks and trains;
 - d. reforming the structure of credits and taxes to remove misaligned incentives;
 - e. reforming federal and state subsidies to incentivize clean technology infrastructure.
8. To transform our energy infrastructure, we must realign federal and state rules, regulations, and rewards to stimulate manufacturing of and markets for clean and efficient energy systems. Such a transformation would be beneficial for our health, for the environment, for sustained economic health, and would contribute to stabilizing the global climate.

Acknowledgments

The authors would like to acknowledge Amy Larkin of Greenpeace, who commissioned Kevin Eckerle, then an independent consultant, to perform work similar to this that is currently unpublished, and subsequently gave permission to make use of their work for this report. We would also like to thank James Hansen, Mark Jacobson, Jonathan Levy, John Evans, and Joel Schwartz for their helpful comments

throughout the course of this work. The genesis for this paper was a Conference—"The True Costs of Coal: Building a Healthy Energy Future"—held October 15–16, 2009 in Washington, DC, supported by the Energy Foundation and the Rockefeller Family Fund.

Conflicts of interest

The authors declare no conflicts of interest.

References

1. International Energy Agency. 2007. *World Energy Outlook 2007: China and India Insights*. International Energy Agency. Paris, France.
2. Energy Information Administration. 2010. U.S. Energy Information Administration. www.eia.doe.gov (accessed December 9, 2010).
3. Energy Information Administration. 2009. *Annual Energy Outlook 2009, with Projections to 2030*. 227 pp. U.S. Department of Energy. Washington, DC.
4. House, K.Z., C.F. Harvey, M.J. Aziz & D.P. Schrag. 2009. The energy penalty of post-combustion CO₂ capture & storage and its implications for retrofitting the U.S. installed base. *Energy & Environmental Science* 2: 193.
5. Katzer, J., E.J. Moniz, J. Deutch, *et al.* 2007. The future of coal: an interdisciplinary MIT study. Technical report, Massachusetts Institute of Technology, Cambridge, MA.
6. Energy Information Administration. 2010. *Energy Information's Outlook Through 2035*. From the Annual Energy Outlook 2010, Diane Kearney, Energy Information Administration, U.S. Department of Energy, Surface Transportation Board, March 23, 2010, Washington, DC.
7. Energy Information Administration. 2008. *Annual Coal Report*. http://www.eia.doe.gov/cneaf/coal/page/acr/acr_sum.html (accessed December 9, 2010).
8. Glustrom, L. 2009. Coal: Cheap and Abundant . . . Or Is It? http://www.cleanenergyaction.org/sites/default/files/Coal_Supply_Constraints_CEA_021209.pdf (accessed December 9, 2010).
9. Heinberg, R. 2009. *Blackout: Coal, Climate and the Last Energy Crisis*. Clairview Books. East Sussex, UK. 208pp.
10. Luppens, J. A. *et al.* 2008. *Assessment of Coal Geology, Resources, and Reserves in the Gillette Coalfield, Powder River Basin, Wyoming: U.S. Geological Survey Open-File Report 2008-1202*.
11. Ruppert, L.F., M.A. Kirschbaum, P.D. Warwick, *et al.* 2002. The US Geological Survey's national coal resource assessment: the results. *Internat. J. Coal Geo.* 50: 247–274.
12. Patzek, T.W. & G.D. Croft. 2010. A global coal production forecast with multi-Hubbert cycle analysis. *Energy* 35: 3109–3122.
13. Krauss, C. 2008. An export in solid supply. *New York Times*.
14. Goodell, J. 2006. *Big Coal: The Dirty Secret Behind America's Energy Future*. Houghton Mifflin. NY.

15. Ghose, M. 2007. Generation and quantification of hazardous dusts from coal mining in the Indian context. *Environ. Monit. Assess.* **130**: 35–45.
16. Palmer, M.A., E.S. Bernhardt, W.H. Schelsinger, *et al.* 2010. Science and regulation. Mountaintop mining consequences. *Science* **327**: 148–149.
17. Hendryx, M. 2009. Mortality from heart, respiratory, and kidney disease in coal mining areas of Appalachia. *Int. Arch. Occup. Environ. Health* **82**: 243–249.
18. Hendryx, M. & M. Ahern. 2008. Relations between health indicators and residential proximity to coal mining in West Virginia. *Am. J. of Public Health* **98**: 669–671.
19. Hendryx, M. & M. Ahern. 2009. Mortality in Appalachian coal mining regions: the value of statistical life lost. *Public Health Rep.* **124**: 541–550.
20. National Research Council. 2009. *The Hidden Costs of Energy: Unpriced Consequences of Energy Production*. Washington, DC.
21. Levy, J., L. Baxter & J. Schwartz. 2009. Uncertainty and variability in health-related damages from coal-fired power plants in the United States. *Risk Analysis* **29**: 1000–1014.
22. Trasande, L., P. Landrigan & C. Schechter. 2005. Public health and economic consequences of methyl mercury toxicity to the developing brain. *Environ. Health Perspect.* **113**: 590–596.
23. Trasande, L., C. Schechter, K. Haynes & P. Landrigan. 2006. Mental retardation and prenatal methylmercury toxicity. *Am. J. Ind. Med.* **49**: 153–158.
24. Epstein, P.R. & J. Selber. 2002. *Oil: A Life Cycle Analysis of its Health and Environmental Impacts*. The Center for Health and the Global Environment. Harvard Medical School. Boston, MA.
25. Pré Consultants. 2008. SimaPro 7.1. <http://www.pre.nl/> (accessed December 9, 2010).
26. Bickel, P. & R. Friedrich. Eds. 2004. *ExternE: Externalities of Energy. Methodology 2005 Update*. European Commission. Luxembourg.
27. Mine Safety and Health Administration. 2008. <http://www.msha.gov/> (accessed December 9, 2010).
28. Yardley, J. 2008. As most of China celebrates New Years, a scramble continues in coal country. *New York Times*, February 9.
29. Jacobs, A. 2010. Rescuers in China struggle to free 153 trapped miners. *New York Times*, March 29.
30. National Institute for Occupational Safety and Health. 2008. What's New in the CWHSP. *NIOSH Coal Worker's Health Surveillance Program*. <http://www.cdc.gov/niosh/topics/surveillance/orads/pdfs/CWHSP-News-Fall2008.pdf> (accessed December 9, 2010).
31. Lloyd, J. W. 1971. Long-term mortality study of steelworkers V. Respiratory Cancer in coke plant workers. *J. Occ. Med.* **12**: 151–157.
32. Zeller, T.A. 2010. A battle in mining country pits coal against wind. *New York Times*, August 14.
33. U.S. Environmental Protection Agency. 2010. EPA Makes Announcement on Two Proposed West Virginia Mountaintop Coal Mines. *News Release*. <http://yosemite.epa.gov/opa/advpress.nsf/bd4379a92ceceac8525735900400c27/84636183a97ced24852576a20069961a!OpenDocument> (accessed December 9, 2010).
34. Kentucky Division of Water. 2008. Integrated Report to Congress on Water Quality in Kentucky. <http://www.water.ky.gov/sw/swmonitor/305b/> (accessed December 9, 2010).
35. Office of Surface Mining. 2008. *Final Environmental Impact Statement: Excess Spoil Minimization, Stream Buffer Zones*. Washington, DC. <http://www.osmre.gov/aml/AML.shtm> (accessed December 17, 2010).
36. U.S. Environmental Protection Agency. 2002. *Affected Environment and Consequences of Mountain-top Removal and Valley Fill Practices. Environmental Impact Statement*. Washington, DC <http://www.epa.gov/region3/mnttop/pdf/III.affected-envt-consequences.pdf> (accessed December 17, 2010).
37. Epstein, P.R., W. Moomaw & C. Walker. 2008. Healthy solutions for the low carbon economy: guidelines for investors, insurers, and policy makers. The Center for Health and the Global Environment. Harvard Medical School. Boston, MA.
38. Fox, J.F. & J.E. Campbell. 2010. Terrestrial carbon disturbance from mountaintop mining increases lifecycle emissions for clean coal. *Environ. Sci. Technol.* **44**: 2144–2149.
39. Groisman, P.Y., R.W. Knight, T.R. Karl, *et al.* 2004. Contemporary changes of the hydrological cycle over the contiguous United States: trends derived from in situ observations. *Journal of Hydrometeorology* **5**: 64–85.
40. Groisman, P.Y. & R.W. Knight. 2008. Prolonged dry episodes over the conterminous United States: new tendencies emerging during the last 40 years. *BAMS* **21**: 1850–1862.
41. Reinhart, N. 2010. Personal communication.
42. Intergovernmental Panel on Climate Change, United Nations Environment Programme. Technology and Economics Assessment Panel. 2005. IPCC special report on carbon dioxide capture and storage: summary for policymakers. 25 pp. Geneva, Switzerland. http://www.ipcc.ch/pdf/special-reports/srccs/srccs_wholereport.pdf (accessed December 17, 2010).
43. Forster, P., V. Ramaswamy, P. Artaxo, *et al.* 2007. Changes in Atmospheric Constituents and in Radiative Forcing. *Climate Change 2007: The Physical Science Basis. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change* 1–106.
44. West Virginia Department of Environmental Protection. 2010. <http://www.dep.wv.gov/Pages/default.aspx> (accessed December 9, 2010).
45. U.S. Environmental Protection Agency. 2010. EPA Facilities Registry. http://www.epa.gov/enviro/html/fii/fii_query.java.html (accessed December 9, 2010).
46. Dewan, S. 2009. Hundreds of coal ash dumps lack regulation. *New York Times*, January 6.
47. Environmental Integrity Project. 2009. Coming Clean: What the EPA Knows About the Dangers of Coal Ash. A Summary of the United States Environmental Protection Agency's 2007 Human and Ecological Risk

- Assessment of Coal Combustion Wastes. http://www.environmentalintegrity.org/news_reports/news_09_05_07.php (accessed December 9, 2010).
48. Stout, B.M. & M.E. Cassidy. 2009. Personal communication.
 49. Wheeling Jesuit University. 2006. Kentucky Coal Impoundment Location and Information System. <http://www.coalimpoundment.org/locate/list.asp> (accessed December 9, 2010).
 50. Spath, P., M. Mann & D. Kerr. 1999. Life cycle assessment of coal-fired power production. Dept. of Energy, Oak Ridge, TN. <http://www.nrel.gov/docs/fy99osti/25119.pdf> (accessed December 9, 2010).
 51. Stout, B.M. & M.E. Cassidy. 2006. Environmental and social-economic impacts of electrical generation from the fossil fuel cycle in the Mid-Atlantic Highlands (unpublished). 1–25.
 52. U.S. Department of Energy, U.S. Department of Health and Human Services, U.S. Environmental Protection Agency. 1980. *Health effects of coal technologies: research needs*. Washington, DC. <http://mdl.csa.com/partners/viewrecord.php?requester=gs&collection=TRD&recid=20070430115163CE&q=Health+effects+of+coal+technologies%3A+research+needs.&uid=790182396&setcookie=yes> (accessed December 17, 2010).
 53. U.S. Department of Energy. 1986. *Effects of Coal Rank and the Chemical Composition and Toxicological Activity of Coal Liquefaction Materials*. Washington, DC.
 54. U.S. Environmental Protection Agency. 2009. Clean Air Mercury Rule. <http://www.epa.gov/mercuryrule/basic.htm> (accessed December 9, 2010).
 55. Kentucky Department of Fish and Wildlife Resources. 2009. Fish Consumption Advisories. <http://www.fw.ky.gov/fishadvisory.asp> (accessed December 9, 2010).
 56. Clean Air Task Force. Power Plant Pollution Locator. 2010. http://www.catf.us/projects/power_sector/power_plant_emissions/pollution_locator/ (accessed December 9, 2010).
 57. U.S. Environmental Protection Agency Clean Air Markets Division. 2004. Memorandum: Economic and energy analysis of the proposed interstate air quality rulemaking. <http://www.epa.gov/cair/pdfs/tm0009.pdf> (accessed December 17, 2010).
 58. Keating, M., L. Evans, B. Dunham & J. Stant. 2009. Waste Deep: Filling Mines with Coal Ash is Profit for Industry, But Poison for People. <http://www.earthjustice.org/news/press/2009/new-report-documents-unseen-threat-to-toxic-coal-ash> (accessed December 9, 2010).
 59. Koornneef, J., T. Van Keulen, A. Faaij & W. Turkenburg. 2008. Life cycle assessment of a pulverized coal power plant with post-combustion capture, transport and storage of CO₂. *Int. J.Greenhouse Gas Control* **2**: 448–467.
 60. Hendryx, M., K. O'Donnell & K. Horn. 2008. Lung cancer mortality is elevated in coal-mining areas of Appalachia. *Lung Cancer* **62**: 1–7.
 61. Shiber, J. 2005. Arsenic in domestic well water and health in central Appalachia, USA. *Water, Air, Soil Pollut.* **160**: 327–341.
 62. Hendryx, M. 2008. Mortality rates in Appalachian coal mining counties: 24 years behind the nation. *Environ. Justice* **1**: 5–11.
 63. Ahern, M., M. Mullett, K. MacKay & C. Hamilton. 2010. Residence in coal-mining areas and low-birth-weight-outcomes. *Maternal Child Health J.* Epub ahead of print. doi: 10.1007/s10995-009-0555-1/ <http://www.ncbi.nlm.nih.gov/pubmed/20091110> (accessed December 9, 2010).
 64. Ahern, M., K. MacKay, D. Carpenter & M. Hendryx. Low birth weight and pre-term birth outcomes in coal mining areas of the United States. *J. Environ. Health* DOI: 10.1007/s10995-009-0555-1/ <http://www.ncbi.nlm.nih.gov/pubmed/20091110> (accessed December 17, 2010).
 65. Tyson, J.E. & S. Saigal. 2005. Outcomes for extremely low-birth-weight infants: disappointing news. *JAMA* **294**: 371–373.
 66. Curhan, G. C. *et al.* 1996. Birth weight and adult hypertension, diabetes mellitus, and obesity in U.S. men. *Circulation* **94**: 3426–3250.
 67. Hendryx, M. & M. Ahern. 2010. Appalachian Coal Mining & Human Health. *Spring 2010 Symposium, Mountaintop Coal Mining: Human Health and Ecological Concerns*. Presentation available upon request.
 68. Hendryx, M. 2010. Personal communication.
 69. Morse, J.C., B.P. Stark & W.P. McCafferty. 1993. Southern Appalachian streams at risk: implications for mayflies, stoneflies, caddisflies, and other aquatic biota. *Aqua. Conserv.: Marine Freshwater Ecosys.* **3**: 293–303.
 70. Bragg, E.A. 1999. Civil Action No. 2:98–0636. United States District Court Southern District of West Virginia Civil Action No. 2:98–0636. United States District Court Southern District of West Virginia. <http://www.publicjustice.net/Repository/Files/Bragg%20district%20court%20PI%20decision.pdf> (accessed December 9, 2010).
 71. Stout, B. & B. Wallace. 2005. A Survey of Eight Major Aquatic Insect Orders Associated with Small Headwater Streams Subject to Valley Fills from Mountaintop Mining. 1–13. <http://www.epa.gov/region3/mtntop/pdf/appendices/d/StoutWallaceMacroinvertebrate.pdf> (accessed December 17, 2010).
 72. Gilbert, N. 2010. Mountain mining damages streams. *Nature* **466**: 806.
 73. Kirk, E.J., R.A. Johnston & R. Maggard. 2010. An evaluation of the mayfly abundances and WV-SCI scores compared to levels of conductivity in several streams in southern West Virginia. *2010 West Virginia Surface Mine Drainage Task Force Symposium Papers*, Morgantown, WV, March 30–31. Conference: <http://wvmdtaskforce.com/proceedings/2010.cfm>. Paper: <http://wvmdtaskforce.com/proceedings/10/Kirk-REIC-Argus-Mayfly-Conductivity-Paper.doc> (accessed December 17, 2010).
 74. Pond, G.J. 2010. Patterns of Ephemeroptera taxa loss in Appalachian headwater streams (Kentucky, USA). *Hydrobiologica* <http://www.springerlink.com/content/b1w1153363782687/> (accessed December 17, 2010).

75. Pond, G.J., M. E. Passmore, F.A. Borsuk, *et al.* 2008. Downstream effects of mountaintop coal mining: comparing biological conditions using family- and genus-level macroinvertebrate bioassessment tools. *J. N. Am. Benthol. Soc.* **27**: 717–737.
76. Anon. 2007. Conservation: mines vs. mussels. *Nat. Geo. Magazine* **212**.
77. Merritt, R.W. & K.W. Cummins. 1996. *An Introduction to the Aquatic Insects of North America*. Kendall/Hunt Publishing Company. Dubuque, IA.
78. Palmer, M.A., E.S. Bernhardt, E.A. Chornesky, *et al.* 2005. Ecological science and sustainability for the 21st century. *Frontiers Ecol. Environ.* **3**: 4–11.
79. Hendryx, M. & K.J. Zullig. 2009. Higher coronary heart disease and heart attack morbidity in Appalachian coal mining regions. *Preven. Med.* **49**: 355–359.
80. U.S. Census Bureau. 2000. Net migration for the population 5 years and over for the United States, regions, states, counties, New England minor civil divisions, metropolitan areas, and Puerto Rico. Washington, DC.
81. U.S. Environmental Protection Agency. 2009. Emissions & Generation Resource Integrated Database. <http://www.epa.gov/cleanenergy/energy-resources/egrid/index.html> (accessed December 9, 2010).
82. U.S. Environmental Protection Agency. 2009 National Emissions Inventory. <http://www.epa.gov/oar/data/neidb.html> (accessed December 9, 2010).
83. Pope, C.A. & D.W. Dockery. 2006. Health effects of fine particulate air pollution: lines that connect. *J. Air & Waste Manage. Assoc.* **56**: 709–742.
84. Levy, J. & J. Spengler. 2002. Modeling the benefits of power plant emission controls in Massachusetts. *J. Air & Waste Manage. Assoc.* **52**: 5–18.
85. Schneider, C. & M. Padian. 2004. *Dirty Air, Dirty Power: Mortality and Health Damage Due to Air Pollution from Power Plants*. Clean Air Task Force. Boston, MA.
86. Levy, J., J. Hammitt & J. Spengler. 2000. Estimating the mortality impacts of particulate matter: What can be learned from between-study variability? *Environ. Health Perspect.* **108**: 109–117.
87. O'Neill, M.S., A. Veves, A. Zanobetti, *et al.* 2005. Diabetes enhances vulnerability to particulate air pollution-associated impairment in vascular reactivity and endothelial function. *Circulation* **111**: 2913–2920.
88. Ritz, B., M. Wilhelm & Y. Zhao. 2006. Air pollution and infant death in southern California, 1989–2000. *Pediatrics* **118**: 493–502.
89. Levy, J.I., J.D. Spengler, D. Hlinka & D. Sullivan. 2000. *Estimated Public Health Impacts of Criteria Pollutant Air Emissions from the Salem Harbor and Brayton Point Power Plants*. Harvard School of Public Health. Boston, MA.
90. Schwartz, J., B. Coull, F. Laden & L. Ryan. 2007. The effect of dose and timing of dose on the association between airborne particles and survival. *Environ. Health Perspect.* **116**: 64–69.
91. Schwartz, J., 2010. Personal communication.
92. U.S. Environmental Protection Agency. 2009. National Ambient Air Quality Standards. <http://www.epa.gov/air/criteria.html> (accessed December 9, 2010).
93. Dockery, D.W. *et al.* 1993. An association between air pollution and mortality in six U.S. cities. *Engl. J. Med.* **329**: 1753–1759.
94. Roman, H., K.D. Walker, T.L. Walsh, *et al.* 2008. Expert judgment assessment of the mortality impact of changes in ambient fine particulate matter in the US. *Environ. Sci. Technol.* **42**: 2268–2274.
95. Jaworski, N., R. Howarth & L. Hetling. 1997. Atmospheric deposition of nitrogen oxides onto the landscape contributes to coastal eutrophication in the northeast United States. *Environ. Sci. Tech.* **31**: 1995–2004.
96. Townsend, A.R., W. Howarth, F.A. Bazzaz, *et al.* 2003. Human health effects of a changing global nitrogen cycle. *Frontiers Ecol. Environ.* **1**: 240–246.
97. Heisler, J. *et al.* 2008. Eutrophication and harmful algal blooms: a scientific consensus. *Harmful Algae* **8**: 3–13.
98. Kirkpatrick, B., L.E. Fleming, D. Squicciarini, *et al.* 2004. Literature review of Florida red tide: implications for human health effects. *Harmful Algae* **3**: 99–115.
99. Shumway, S. 1990. A review of the effects of algal blooms on shellfish and aquaculture. *J. World Aquacul. Soc.* **21**: 65–104.
100. Wang, J. & J. Wu. 2009. Occurrence and potential risks of harmful algal blooms in the East China Sea. *Sci. Total Environ.* **407**: 4012–4021.
101. National Centers for Ocean Coastal Science. 2009. Economic Impacts of Harmful Algal Blooms. http://www.cop.noaa.gov/stressors/extremeevents/hab/current/econimpact_08.pdf (accessed December 9, 2010).
102. Hoagland, P. & S. Scatista. 2006. *The Economic Effects of Harmful Algal Blooms*. Springer-Verlag. Dordrecht, the Netherlands.
103. Carlisle, E. 2009. The Gulf of Mexico Dead Zone and Red Tides. <http://www.tulane.edu/~bflcury/envirobio/enviroweb/DeadZone.htm> (accessed December 9, 2010).
104. Driscoll, C.T., G.B. Lawrence, A.J. Bulger, *et al.* 2001. Acid rain revisited: advances in scientific understanding since the passage of the 1970 and 1990 Clean Air Act Amendments. Hubbard Brook Research Foundation. Science Links' Publication. Vol. 1, no.1, Hanover, NH.
105. Yanai, R.D., T.G. Siccama, M.A. Arthur, *et al.* 1999. Accumulation and depletion of base cations in forest floors in the northeastern US. *Ecology* **80**: 2774–2787.
106. Banzhaf, H., D. Burtraw, D. Evans & A. Krupnick. 2004. Valuation of natural resource improvements in the Adirondacks. Resources for the Future, Washington, DC. 40 pp <http://www.rff.org/rff/Documents/RFF-RPT-Adirondacks.pdf> (accessed December 17, 2010).
107. Northeast States for Coordinated Air Use Management. 2003. Mercury emissions from coal-fired power plants: the case for regulatory action. NESCAUM, Boston, MA. www.nescaum.org/documents/rpt031104mercury.pdf/ (accessed December 9, 2010).
108. Rice, G.E., J.K. Hammitt & J.S. Evans. 2010. A probabilistic characterization of the health benefits of reducing methyl mercury intake in the United States. *Environ. Sci. Tech.* **44**: 5216–5224.

109. IPCC. 2007. Climate Change 2007: Synthesis Report. Contribution of Work Groups I, II, and III to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change. IPCC. Geneva, Switzerland.
110. U.S. Environmental Protection Agency. 2008. Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990–2006. Washington, DC.
111. Hansen, J., M. Sato, P. Kharecha, *et al.* 2007. Climate change and trace gases. *Philos. Transact. A Math Phys Eng Sci.* **365**: 1925–1954.
112. Jacobson, M.Z. 2002. Control of fossil-fuel particulate black carbon and organic matter, possibly the most effective method of slowing global warming. *J. Geophys. Res.* **107**: (D19), 4410, doi:10.1029/2001JD001376, 22pp. [pp ACH 16-1 to 16-22]. <http://www.stanford.edu/group/efmh/fossil/fossil.pdf> (accessed December 9, 2010).
113. Ramanathan, V. & G. Carmichael. 2008. Global and regional climate changes due to black carbon. *Nat. Geosci.* **1**: 221–227.
114. Levitus, S., J. Antonov & T. Boyer. 2005. Warming of the world ocean, 1955–2003. *Geophys. Res. Lett.* **32**: L02604, doi:10.1029/2004GL021592.
115. Barnett, T., D.W. Pierce, K.M. AchutaRao, *et al.* 2005. Penetration of human-induced warming into the world's oceans. *Science* **309**: 284–287.
116. Epstein, P.R. & J.J. McCarthy. 2004. Assessing climate stability. *Bull. Am. Meteorological Soc.* **85**: 1863–1870.
117. IPCC. 2007. Summary for Policymakers. In: *Climate Change 2007: Impacts, Adaptation, and Vulnerability. Contribution of Working Group II to the Fourth Assessment Report of the Intergovernmental Panel of Climate Change*. Cambridge University Press. Cambridge, UK. 7–22.
118. Cooke, W., C. Lioussé, H. Cachier & J. Feichter. 1999. Construction of a 1 × 1 fossil fuel emission data set for carbonaceous aerosol and implementation and radiative impact in the ECHAM4 model. *J. Geophys. Res.* **104**: 22137–22162.
119. P.R. Epstein, P.R. & E. Mills (Eds.) 2005. *Climate Change Futures: Health, Ecological and Economic Dimensions, Center for Health and the Global Environment*, Harvard Medical School, Boston, MA.
120. Thompson, C. 2010. Disaster capitalism. *Mother Jones* July/Aug. 2010.
121. Stern, N. 2003. *Stern Review on the Economics of Climate Change*. Cambridge University Press. Cambridge, UK. www.nescaum.org/documents/rpt031104mercury.pdf (accessed December 17, 2010).
122. Haszeldine, R. 2009. Carbon capture and storage: how green can black be? *Science* **325**: 1647–1652.
123. General Accounting Office. 2008. *Climate Change: Federal Actions Will Greatly Affect the Viability of Carbon Capture and Storage as a Key Mitigation Option*. Washington, DC.
124. Renard, F., E. Gundersen, R. Hellmann, *et al.* 2005. Numerical modeling of the effect of carbon dioxide sequestration on the rate of pressure solution creep in limestone: Preliminary results. *Oil & Gas Sci. Tech.* **60**: 381–399.
125. KNBC. 2006. Coroner: carbon dioxide killed men on mammoth mountain. <http://www.knbc.com/news/8516441/detail.html> (accessed December 9, 2010).
126. Konty, M.F. & J. Bailey. 2009. The impact of coal on the Kentucky State Budget. <http://www.maced.org/coal/> (accessed December 9, 2010).
127. Energy Information Administration. 2008. Federal Financial Interventions and Subsidies in Energy Markets 2007. <http://www.eia.doe.gov/oiaf/servicerpt/subsidy2/index.html> (accessed December 9, 2010).
128. Environmental Law Institute. 2009. Estimating U.S. Government subsidies to energy sources: 2002–2008. 1–37. Washington, DC. http://www.eli.org/Program_Areas/innovation_governance.energy.cfm (accessed December 17, 2010).
129. Office of Surface Mining Reclamation and Enforcement. 2009. Estimating U.S. Government Subsidies to Energy Sources: 2002–2008. <http://www.osmre.gov/aml/AML.shtm> (accessed December 9, 2010).
130. Lockie, S., M. Franetovich, V. Petkova-Timmer, *et al.* 2009. Coal mining and the resource community cycle: a longitudinal assessment of the social impacts of the Coppabella coal mine. *Environ. Impact Asses.* **29**: 330–339.
131. Wood, L.E. 2005. Trends in national and regional economic distress: 1960–2000. <http://www.arc.gov/assets/research/reports/TrendsInNationalandRegionalEconomicDistress1960to2000.pdf> (accessed December 17, 2010).
132. Marmot M. & R.G. Wilkinson. Eds. 2006. *Social Determinants of Health*. 2nd ed. Oxford University Press. Oxford, UK.